



RADIOLOGICAL EVALUATION OF BUILDING MATERIALS USED IN MALUMFASHI, KATSINA STATE, USING GAMMA-RAY SPECTROSCOPY ANALYSIS

¹Aku, M.O. and ²Yusuf, U.

¹Physics Department, Bayero University, Kano, Kano, State Nigeria

²Physics Department, SBRS Funtua, Ahmadu Bello University, Zaria, Kaduna State Nigeria

ABSTRACT

The activity concentrations of naturally occurring radionuclides (^{226}Ra , ^{232}Th and ^{40}K) in building materials (sand, cement, blocks, granite, and paints) used in the construction of buildings in Malumfashi local Government area of Katsina state, Nigeria were determined by means of a gamma-ray spectrometry system using Sodium Iodide thallium activated (NaI(Tl)) detector in a low background configuration. The average activity concentrations for ^{226}Ra , ^{232}Th and ^{40}K ranged from 17.07 ± 4.98 to 50.72 ± 6.74 , 23.03 ± 1.54 to 42.91 ± 2.37 and 84.03 ± 5.44 to $363.45 \pm 5.27 \text{ Bqkg}^{-1}$ respectively. Except for block samples, all the samples mean activity concentrations were less than their respective world averages. The estimated absorbed gamma dose rate (D) ranged from 30.75 ± 4.22 to $64.57 \pm 4.78 \text{ nGyh}^{-1}$ and the outdoor annual effective dose equivalent (AEDE) varied from 0.038 ± 0.004 to $0.06 \pm 0.0051 \text{ mSvy}^{-1}$; these values are below the worldwide effective dose of 0.07 mSvy^{-1} . The values of radium equivalent activity (R_{eq}), the external hazard index (H_{ex}) and the internal hazard index (H_{in}) for all the samples in the present work were lower than the accepted safety limit value of 370 Bqkg^{-1} .

Keywords: Radionuclide, hazard indices, gamma-ray spectrometry, building materials, Malumfashi

INTRODUCTION

All building raw materials and products derived from rock and soil contain various amounts of mainly natural radionuclides of the uranium (^{238}U) and thorium (^{232}Th) series, and the radioactive isotope of potassium (^{40}K) (El-Taher, 2010). In the ^{238}U -series the decay chain segment starting from radium (^{226}Ra) is radiologically the most important and therefore, reference is often made to ^{226}Ra instead of ^{238}U (Al-Jundi *et al.*, 2009). These radionuclides are sources of the external and internal radiation exposures in dwellings. The external exposures is usually a result of direct gamma radiation while the inhalation of radioactive inert gases radon (^{222}Rn , a daughter product of ^{226}Ra) and their short lived secondary product lead to the internal exposure of the respiratory tract to alpha particles. The specific activities of ^{226}Ra , ^{232}Th and ^{40}K in building materials and products mainly depend on geological and geographical condition as well as geochemical characteristics of those materials (UNSCEAR, 1993 and El-Taher, 2007).

The radiological impact of the natural radioactivity is due to radiation exposure of the body by gamma-rays and irradiation of lung tissues from inhalation of radon and its progeny. From the natural risk point of view, it is necessary to know the dose limit of public exposure and to measure the natural radiation level in the environment caused by ground, air, water, food, building interiors, etc. (El-Taher and Makhluaf, 2010). Low level gamma-ray spectrometry is suitable for both qualitative and quantitative determination of gamma-

ray emitting nuclides in the environment. The concentration of radioactive element in building materials and its components are important in assessing population exposures, as most individuals spend 80% of their time indoors. The average indoor absorbed dose rate in air from terrestrial sources of radioactivity is estimated to be 70 nGyh^{-1} (El-Taher, 2012). Great attention has been paid to determining radionuclide's concentration in building materials in many countries (Al-jundi *et al.*, 2009). In building, the highest concentration of radionuclides are found in mineral based materials such as stone, sand, bricks, cement and sediment (Jonathan *et al.*, 2013). The present work therefore is devoted to determine the presence of radioactive elements (^{226}Ra , ^{232}Th , and ^{40}K) in building materials used in Malumfashi Local Government area, assess the radiological hazards to human health, measure activity concentrations for these natural radionuclides and calculate the radiological parameters (radium equivalent activity R_{eq} , external hazard index H_{ex} , internal hazard index H_{in} , absorbed dose rate D and the annual effective dose equivalent AEDE). These radiological parameters will be compared with reported values for other countries in order to determine the hazard effects. The results will be essential for development of data base on the use and management of building materials. Gamma-ray spectrometry analytical techniques were chosen in this work due to their sensitivity, simplicity and accuracy especially with regard to the analysis of building materials and other environmental samples.

MATERIALS AND METHODS

The analysis of this work was carried out at the Centre for Energy Research and Training (CERT), Ahmadu Bello University, Zaria. Sample of granite, sand, blocks, cement and paint were collected in four different locations namely Malumfashi east, Malumfashi south, Malumfashi west, and Malumfashi north, analyzed and all the data obtained (in cps) were converted to the conventional unit of Bqkg⁻¹ using the calibration factors. Energy and efficiency calibration as well as background measurements are part of analytical parameters required for ease of sample analysis and data interpretation in measurements of naturally occurring radioactive materials (NORMs) with gamma spectrometry method

$$CF_K = \frac{cps(^{40}K)}{Bq(^{40}K)/kg}, \quad CF_{Ra} = \frac{cps(^{226}Ra)}{Bq(^{226}Ra)/kg}$$

A total of twenty one samples were collected and each sample was put into polyethylene bags, tied and labeled appropriately. Double bags were used for each sample to prevent breakage and cross contamination of sample. The samples were air dried for 7 hrs under laboratory temperature (ambient temperature of 27^oC) and average relative humidity of 70% (IAEA, 1989). The samples were then grinded and packed to fill already weighed cylindrical plastic containers of dimension 7.2 cm in diameter and 6.0 cm high to satisfy the selected best sample container higher that will adequately match the detector's dimensions, which allowed for the adequate covering of the detector shield (Ibeanu, 1999). Before sealing, the mass of each sample was determined and recorded. The sealing of the sample containers were done in three stages, namely Vaseline wax , candle wax and adhesive masking tape, all done to prevent the escape of ²²²Rn gas, a daughter product of ²²⁶Ra.

$$Ra_{eq} = C_{Ra} + 1.43C_{Th} + 0.077C_K \tag{1}$$

where C_{Ra}, C_{Th} and C_K are the activity concentrations in building materials samples in Bqkg⁻¹ of ²²⁶Ra, ²³²Th and ⁴⁰K respectively. The Ra_{eq} is related to the external gamma-dose and internal dose due to radon and its daughters.

The gamma outdoor absorbed dose rate (D) at 1 m above the ground was calculated using the specific activities of ⁴⁰K, ²²⁶Ra and ²³²Th. The conversion factor used to calculate the absorbed dose rates (UNSCEAR, 1993) is given as:

$$D(nGyh^{-1}) = 0.461C_{Ra} + 0.623C_{Th} + 0.041C_K \tag{2}$$

where C_{Ra}, C_{Th} and C_K, are the specific activities of ²²⁶Ra, ²³²Th and ⁴⁰K in Bqkg⁻¹ respectively.

The annual effective dose equivalent received outdoor by a member was calculated from the absorbed dose rate by applying dose conversion factor of 0.7 SvGy⁻¹ and the occupancy factor for outdoor and indoor was 0.2(5/24) and 0.8(19/24) respectively .The effective dose rates in units of mSvyr⁻¹ are calculated by the following formulae (UNSCEAR, 1993)

$$AEDE (outdoor) (mSvy^{-1}) = D (nGyh^{-1}) \times 8760h \times 0.7SvGy^{-1} \times 0.2 \times 10^{-6} \tag{3}$$

$$AEDE (indoor) (mSvy^{-1}) = D (nGyh^{-1}) \times 8760h \times 0.7 (SvGy^{-1}) \times 0.8 \times 10^{-6} \tag{4}$$

and were carried out in accordance with the IAEA recommendations (IAEA, 1989).

The assessment of the activity concentration of ²²⁶Ra, ²³²Th and ⁴⁰K were done using 1764keV γ-line of ²¹⁴Bi, 2614keV γ-line of ²⁰⁸Tl and 1460keV γ-line of ⁴⁰K respectively. The net peak count rate (cps) for the radionuclide in any sample was obtained by subtracting net peak count rate of the radionuclide from that of the sample. All the obtained cps were converted to conventional units using the calibration factors CF_K, CF_{Ra} and CF_{Th} derived by Ibeanu (1999) to determine the activity concentration of radionuclide of interest. The calibration factor values of 6.431 x 10⁻⁴, 8.632 x 10⁻⁴ and 8.768 x 10⁻⁴ in cps/Bqkg⁻¹ for CF_K, CF_{Ra} and CF_{Th} respectively were obtained using the following:

$$and \quad CF_{Th} = \frac{cps(^{232}Th)}{Bq(^{232}Th)/kg}$$

The samples were then stored for a minimum period of four weeks to achieve an approximate secular equilibrium between ²²²Rn and ²³²Th, and their respective progeny before commencement of measurement. The samples were then counted for 29,000 sec in a low-level gamma counting spectrometer comprising a 7.6 cm x 7.6 cm NaI (TI) detector which is coupled to multichannel analyzer (MCA) through a preamplifier base. The spectral and live times of the NORMs were acquired using MAESTRO software.

Radium equivalent activity (Ra_{eq}) is an index that has been introduced to represent the specific activities of ²²⁶Ra, ²³²Th and ⁴⁰K by a single quantity, which takes into account the radiation hazards associated with them. This first index can be calculated according to Nuclear Energy Agency – Organization for Economic cooperation and Development (NEA-OECD, 1979) as:

The external hazard index (H_{ex}) can be defined as (OECD 1979)

$$H_{ex} = \frac{C_K}{4810Bqkg^{-1}} + \frac{C_{Ra}}{370Bqkg^{-1}} + \frac{11 C_{Th}}{259Bqkg^{-1}} \leq 1 \quad (5)$$

where C_K , C_{Ra} and C_{Th} are the specific activities of ^{226}Ra , ^{232}Th and ^{40}K in $Bqkg^{-1}$ respectively
 The internal hazard index (H_{in}) can be defined as (OECD 1979).

$$H_{in} = \frac{C_K}{4810Bqkg^{-1}} + \frac{C_{Ra}}{180Bqkg^{-1}} + \frac{C_{Th}}{259Bqkg^{-1}} \leq 1 \quad (6)$$

where C_{Ra} , C_{Th} and C_K are activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K , respectively in $Bqkg^{-1}$. These indices must be less than unity in order to keep the radiation hazard insignificant for the people living in the investigated area (UNSCEAR, 1993).

RESULTS AND DISCUSSION

The average measured radioactivity concentrations of (^{226}Ra , ^{232}Th , ^{40}K) are presented in Table1. The mean activity concentrations of ^{226}Ra in sand, granite, cement, paint are lower than that of the world mean and that of block sample is slightly higher than that

of the world mean for soil, $35 Bqkg^{-1}$. Also the average concentration of ^{232}Th in block sample is slightly higher than that of the world mean for soil $30 Bqkg^{-1}$ (UNSCEAR, 2000). While the activity levels of all the samples are within the world wide range.

Table 1: Average mean activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K for building materials used in study area and that of World Mean.

S/NO	MATERIALS	C_K ($Bqkg^{-1}$)	C_{Ra} ($Bqkg^{-1}$)	C_{Th} ($Bqkg^{-1}$)
1	Cement	210.00±4.19	30.25±2.47	29.68±2.97
2	Sand	282.19±6.13	17.39±3.47	36.89±2.09
3	Granite	235.70±2.48	19.41±2.84	23.03±1.54
4	Paint	84.03±5.44	17.07±4.98	25.82±1.52
5	Blocks	363.45±5.27	50.72±6.74	42.91±2.37
6	World Mean	400	35	30
7	World Range	140-850	17-60	11-64

The average values of radium equivalent activity (Ra_{eq}), absorbed dose rate (D), external hazard indices (H_{ex}), internal hazard indices (H_{in}) and annual

effective dose equivalent (AEDE) are presented in Table 2

Table 2: Mean (Ra_{eq}), (D), (H_{ex}), (H_{in}) and (AEDE) for the building materials in Malumfashi

S/N	Materials	Ra_{eq} ($Bqkg^{-1}$)	D($nGyh^{-1}$)	H_{ex}	H_{in}	AEDE($mSvy^{-1}$)
1	Cement	79.77±5.33	36.51±3.07	0.24±0.02	0.33±0.01	0.037±0.003
2	Sand	96.01±7.13	38.80±3.23	0.25±0.03	0.28±0.02	0.043±0.004
3	Blocks	143.87±13.50	64.51±4.78	0.33±0.02	0.52±0.05	0.06±0.005
4	Granite	72.64±5.23	32.56±3.75	0.17±0.01	0.25±0.03	0.056±0.003
5	Paint	60.61±4.50	30.75±4.21	0.16±0.02	0.21±0.04	0.038±0.004
6	world mean	< 370	55	< 1	< 1	0.07

The Ra_{eq} values for the building materials were found to be within the range of $60.61±4.5$ to $143.87±13.50 Bqkg^{-1}$ and are less than $370 Bqkg^{-1}$, which are acceptable for safe use (OECD, 1979). Also the average values of absorbed dose rates (D) are in the range $30.75±4.22$ to $64.51±4.78$ and found to be comparable to the world average of $55 nGyh^{-1}$ (OECD, 1979). The Annual effective dose equivalent (AEDE) for the different building materials samples in this study varied from $0.037 ± 0.003$ to $0.06 ± 0.005$

$mSvy^{-1}$ are below the permissible limit and are comparable with the worldwide effective dose of $0.07 mSvy^{-1}$ (UNSCEAR, 2000). The calculated values of external hazard index obtained were in the range of $0.16 ± 0.02$ to $0.33 ± 0.01$ and the internal hazard index values ranges from $0.21 ± 0.04$ to $0.52 ± 0.05$. The area is underlain by crystalline rocks of the Basement Complex, which greatly influences the relief. The observed variations in concentrations of the radionuclides may be a reflection of the differences in the geochemistry of the samples and

other climatic conditions. The area of study falls within the Tropical Continental Climatic Zone; with a cool dry (harmattan) season from December to February; a hot dry season from March to May; a warm wet season from June to September; a less marked season after rains during

the months of October to November, characterised by decreasing rainfall and a gradual lowering of temperature.

CONCLUSION

The values of calculated radiation parameters such as dose rate, annual effective dose equivalent, radium equivalent, external hazard index and internal hazard

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index are lower than the permissible limit. On the basis of this, none of the samples considered pose any significant radiation hazard and their usage by inhabitants for different kinds of dwellings is considered safe.

RECOMMENDATION

There could be health risks originating from other source(s) and therefore there will be need to carry out similar researches in order to establish a comprehensive and reliable reference data for the building materials, water, soil and other related environmental materials. It is also important to check the risks associated with level of radioactive elements and heavy or toxic elements.