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PREPARATION OF TiO₂-ACTIVATED KAOLINITE COMPOSITE FOR PHOTOCATALYTIC DEGRADATION OF RHODAMINE B DYE

Samsudeen O. Azeez^{1*}, Ismaila O. Saheed¹, Folahan A. Adekola² and Shina S. Salau¹

¹Department of Chemistry and Industrial Chemistry, Faculty of Pure and Applied Sciences, Kwara State University Malete, P.M.B. 1530, Ilorin, 241103, Nigeria
²Department of Industrial Chemistry, University of Ilorin, P.M.B 1515, Ilorin 240003, Nigeria

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ABSTRACT. Rhodamine B (RhB) is one of the prominent dyes discovered to be carcinogenic and mutagenic substance present in the ecosystem. Alteration of the characteristics of TiO₂ through modification with clay influences the behavior of some organic molecules at its surface, and thus its photocatalytic properties. This study investigates the photocatalytic degradation of RhB using TiO₂-activated kaolinite composite (TiO₂-AKC). TiO₂-AKC was prepared in absolute ethanol and characterized using SEM, EDX, TEM, BET and FTIR. The photodegradation of RhB using TiO₂-AKC at room temperature under visible-light was studied at various time intervals and the concentration of RhB was monitored using Ultraviolet-Visible spectrophotometer. Characterisation of TiO₂-AKC revealed an adequate distribution of TiO₂ over the AKC surface, with an enhanced surface area (TiO₂-AKC 808 m²g⁻¹, TiO₂50.0 m²g⁻¹, AKC 20.6 m²g⁻¹) necessary for effective degradation. The degradation was efficient in TiO₂-AKC than in sole TiO₂ resulting in 99.51% and 88.35%, respectively. The experimental data was analysed using zeroth and first order kinetics, parabolic diffusion, and modified Freundlich model. The degradation using TiO₂-AKC fit into modified Freundlich model and parabolic diffusion model while sole TiO₂ was best described by parabolic diffusion model.

KEY WORDS: Kaolin, Kinetics, Organic dyes, Photodegradation, Pollutants, Titanium(IV) oxide

INTRODUCTION

One of the many synthetic dyes is Rhodamine B which had remained a prominent, more useful industrial organic dye. It is used mainly in textile and paper industries, and also applicable in water tracing and cell membrane studies. It is however true that more than 12% of working dyes used in textiles and paper are released as effluents to the environment [1]. These colored effluents usually impact greater degree of environmental pollution, and therefore necessary to treat the effluent before discharging to the environment [2]. Though attempt have been made using several processes such as; adsorption, membrane filtration, chemical coagulation, biological treatment among others for the removal of these toxic materials. However, some of the processes either generate secondary pollutants or were not cost effective. In recent years, photodegradation using semiconductors such as titanium(IV) oxide (TiO₂), as photocatalyst had been employed [3-5].

TiO₂ and ZnO are semiconductors which are often used as catalytic agents due to their high stability, low cost and high efficiency. TiO₂ has been reported to exhibits higher catalytic activity under the ultraviolet radiation than ZnO [1]. Previously, researchers had sought for several ways of degrading dyes. Whereas, degradation of dyes using TiO₂ nanostripe has been applied to dyes degradation in wastewater [6]. Titanium(IV) oxide-activated carbon composite (TiO₂-AC) has also been prepared for degrading Rhodamine B (RhB) dye with high RhB removal efficiency [7].

Kaolinite, a class of layered silicate mineral with one tetrahedral sheet and one octahedral sheet has various uses as a result of its properties such as natural whiteness, fine particle, non-abrasive, and chemical stability [8]. It has been considered a good effluent removal in

^{*}Corresponding author. E-mail: samsudeen.azeez@kwasu.edu.ng, samsudeen.azeez2015@gmail.com This work is licensed under the Creative Commons Attribution 4.0 International License

wastewaters, especially due to its high porous structure, which accounts for its high adsorptive property [9]. Titanium(IV) oxide-kaolinitic clay composite has been considered as a more potent photocatalyst for degradation, due to its high photo stability, cost effectiveness and high degrading efficiency. To the best of our knowledge, there is an unfilled vacuum in research findings into the use of titanium(IV) oxide-activated kolinitic clay composite for photocatalytic degaradation of dyes. Therefore, it is very important to dig-in for alternative routes aiming at more cost effective and environment friendly methods of the previously established degrading types. This research targets investigations into the effectiveness of photocatalytic degradation of Rhodamine B dye from aqueous solution using TiO_2 -AKC composite.

EXPERIMENTAL

Reagents

Analytical grade reagents and chemicals used in this work are as follows: TiO_2 powders (80% anatase, 20% rutile; BET surface area 50 m²/g), Rhodamine B dye (RhB), nitric acid (purity 98%, 1.51g/cm³), and absolute ethanol (BDH Analar, 95% UN No. 1097).

Preparation of sample

The kaolin clay sample was ground, washed with deionised water, dried and sieved into particle size of $<90 \mu$ m. It was impregnated into 1 M HCl for 4 h to remove impurities, washed with deionised water and then dried in the oven at 105 °C for 2 h [10, 11]. A 100 g of the pre-treated sample was activated with 400 mL of 1 M HNO₃ and stirred at 500 rpm for 2½ h at 90 °C. The sample was then filtered, washed to neutral pH with deionized water and oven dried at 105 °C to prepare activated kaolinitic clay (AKC) samples [11].

Composite preparation

A 10 g of AKC was disperse in 1000 mL of deionised water and stirred for 2 h. 7 g of the TiO_2 was mixed with the dispersion to obtain 0.7: 1 TiO_2 -AKC slurry which was then stirred for 24 h. The resulting slurry was centrifuged at 4,000 rpm for 10 min. The solid layer was washed with deionised water and centrifuged. A 1:1 water-ethanolic mixture was then added to the resulting TiO_2 -AKC composite, prior to hydrothermal treatment in an autoclave at 120 °C for 2 h. The composite was re-centrifuged at 4000 rpm for 15 min and oven dried at 110 °C for 1 h. It was then ground to fine particles and stored for further use [12, 13].

Characterization of the prepared composite

Transmission electron microscope (TEM) and scanning electron microscope (SEM) Model JSM-7600F was used for morphological studies. Energy dispersive X-ray spectroscope (EDS), Oxford Instrument was used for chemical composition. Fourier transform infrared spectroscope (FTIR), Model spectrum 100, Perkin-Elmer was employed to determine the functional properties. The surface area and pore dimensions were assessed using Brunauer-Emmett-Teller (BET) instrument.

Photocatalytic experiment

A 0.5 g of TiO₂-AKC composite was taken into 250 mL conical flask and 100 mL of 50 mg/L Rhodamine B was added. Prior to irradiation, the suspension was stirred in the dark for 30 min on a magnetic stirrer for adsorption/desorption equilibria to be established on the surface of the

prepared TiO₂-AKC particles. Thereafter, the suspension was placed under solar light irradiated at $75\pm3x10^3$ I_x and 15 mL of the component mixture were taken at time intervals of 5, 10, 30, 60, 90, 120 min, centrifuged immediately after taken and then filtered through a millipore filter paper (0.22 mm). The residual concentrations of the dye in the solution were analysed using UV-Visible spectrophotometer at 554 nm to determine the extent of degradation of RhB [12, 14]. The percentage of Rhodamine B dye degradation efficiency was calculated using equation 1:

$$D_e(\%) = \frac{Co - Ce}{Co} \times 100 \tag{1}$$

where D_e is the degradation efficiency in percent (%), C_o is the initial concentration of RhB dye prepared, and C_e is the equilibrium concentration of RhB.

Degradation kinetic model

Photodegradation kinetics using TiO₂-AKC composite, under different reaction conditions was examined using zero and first order kinetics. The mechanism of degradation process was monitored using parabolic diffusion and modified Freundlich model [15-16].

Zero order kinetics describes the degradation process and expressed using equation 2,

$$C_e - C_o = -\mathrm{kt} \tag{2}$$

First order kinetics describes the systems where the rate of degradation of dyes depends on the amount of dye molecules and expressed with equation 3,

$$\log\left(\frac{c_e}{c_o}\right) = -\mathrm{kt} \tag{3}$$

Parabolic diffusion model explains the degradation process as diffusion controlled process. The expression is given in equation 4,

$$\frac{1 - \left(\frac{L_{e}}{C_{o}}\right)}{t} = -kt^{-1/2} + a$$
(4)

The modified Freundlich model explains experimental data in terms of ion exchange and diffusion-controlled process. A good fit into the Freundlich model indicates a heterogeneous surface binding. The expression is given in equation 5

$$\log(1 - C_e/C_o) = \log k + b \log t$$
(5)

where, C_e - equilibrium concentration (mg/L), C_o - initial concentration of the dye(mg/L), k - rate constant, t - time, while a and b are non-specific constant obtained from the graphical data.

RESULTS AND DISCUSSION

Scanning electron microscopy (SEM)

Figure 1a shows the SEM micrograph of TiO₂-AKC, pure TiO₂ and AKC. The SEM micrograph of TiO₂-AKC indicates changes in surface morphology with the emergence of spongy discrete particles that differ from pure TiO₂ and AKC. The SEM micrograph of TiO₂-AKC and TiO₂ (Figure 1b) after degradation revealed that the dye molecules were not retained on the pore surfaces, and had been degraded in both TiO₂-AKC and TiO₂ surfaces, respectively [17].



Figure 1a. SEM imaging of (a) TiO_2 -AKC, (b) TiO_2 and (c) AKC.



Figure 1b. SEM images after degradation of (d) TiO_2 -AKC and (e) TiO_2 .



Figure 2. TEM images of (a) TiO_2 -AKC, (b) TiO_2 , (c) after TiO_2 -AKC and (d) after TiO_2 .

Transmission electron microscopy

The TEM micrographs of TiO₂-AKCand TiO₂ before and after photodegradation show that there is internal distribution of TiO₂, over the AKC surface as shown in Figure 2. TiO₂ is uniformly distributed on AKC and the photocatalytic activity of the TiO₂ composite depends on their surface morphology.

Energy dispersed X-ray spectroscopy (EDX) analysis

The EDX analysis was done to ascertain the percentage weight contents in the prepared composite. The spectra as shown in Figure 3 indicate the presence of Ti, Si, O, C and Al has major elements with composition of 15.2%, 8.61%, 48.71%, 22.34%, and 5.15%, respectively. This confirms that the composite preparation was achieved.



Figure 3. EDX spectrum of the TiO₂-AKC composite.



Figure 4a. FTIR spectra TiO2-AKC.



Figure 4b. FTIR spectra of AKC.



Figure 4c. FTIR spectra of TiO₂.



Figure 4d. FTIR spectra of TiO₂-AKC after degradation.

Bull. Chem. Soc. Ethiop. 2022, 36(1)

18

Preparation of TiO₂-activated kaolinite composite for photocatalytic degradation of rhodamine B 19



Figure 4e. FTIR spectra of TiO₂ after degradation.

Fourier transformed infra-red spectroscopy (FTIR) analysis

The FTIR spectra of TiO₂-AKC composite and pure TiO₂ analysis before and after degradation of RhB are presented in Figure 4a-4e. The characteristics absorption signals in AKC spectra at 3694, 3651 and 3619 cm⁻¹ is attributed to OH stretching vibrations; 1032 and 471 cm⁻¹ vibrations are attributed to Si-O [18]. The observed stretching vibration at 3448 and 1648 cm⁻¹ on TiO₂ spectra is assigned to OH stretch and Ti-OH vibration signals respectively [19]. Distinct vibrational signals at 3821 and 3751 cm⁻¹ that appears on TiO₂-AKC composite's spectra could result from the incorporation of TiO₂ onto AKC. Formation TiO₂-AKC composite could also account for the disappearance of distinct OH stretching vibration of TiO₂ at 3446 cm⁻¹ in TiO₂-AKC spectra. Similar characteristics signals were observed on TiO₂-AKC and pure TiO₂ spectra before and after degradation. For instance, the spectra of TiO₂-AKC before and after degradation show slight shift in absorption bands from 3751 cm⁻¹ (before) to 3748 cm⁻¹ (after); 3650 cm⁻¹ (before) to 3653 cm⁻¹ (after). This is an indication that RhB was not retained on the TiO₂-AKC composite and pure TiO₂, and hence a suggestion that photodegradation had taken place at both TiO₂-AKC and pure TiO₂, respectively [20].

Table 1. Brunauer Emmett and Teller (BET) analysis.

Sample	Specific surface area	Average pore volume	Average pore diameter
	(m^2g^{-1})	$(cm^{3}g^{-1})$	(nm)
TiO ₂ -AKC	808	0.274	1.947
Pure TiO ₂	50.0	0.087	9.894
AKC	20.6	0.018	3.48

BET analysis on TiO₂-AKC composite and TiO₂.

Brunauer Emmett and Teller (BET) analysis

Table 1 summarizes BET analysis of the composite. The results reveal that AKC and pure TiO_2 are mesoporous materials. However, the specific surface area and average pore volume of the composite TiO_2 -AKC was found to be microporous. This could be as a result of the processing

method employed for the preparation of the composite. International Union of Pure and Applied Chemistry classified pores as micropore (< 2 nm), mesopores (2-50 nm) and macropores (> 50 nm) [21].

Photocatalytic degradation experiment

Photocatalytic activity experiment of the pure TiO_2 as well as TiO_2 -AKC composites was studied by evaluating the degradation rate of RhB in aqueous solution in the presence of UV light irradiation. As shown in Figure 5a and 5b, the result revealed that the removal of RhB increases gradually with the increasing illumination time. Initially, about 51% and 33% of RhB was adsorbed on TiO₂-AKC and TiO₂, respectively, until adsorption equilibrium was attained at the end of 30 min in the dark. The concentration of RhB decreases with increase in degradation time under light irradiation.



Figure 5a. Degradation of RhB by TiO₂ and TiO₂-AKC composite (C_0 = 50 mg/L, w = 0.5 g, t = 0-120 min, T = 28 ± 2 °C, n = 3, SD = 0.03).

TiO₂-AKC composite shows better performance for degradation of RhB than the pure TiO₂. This could result from the enhancement of photocatalytic degradation activity of pure TiO₂ by AKC, which had been initially lowered during calcination as reported by Tayade *et al.* [22].



Figure 5b.Degradation efficiency of RhB by TiO₂ and TiO₂-AKC composite. (C_0 = 50 mg/L, w = 0.5 g, t = 0-120 min, T = 28 ± 2 °C, n = 3, SD = 0.03).

Degradation kinetic studies

The rate and mechanism of successive adsorption and degradation processes were analysed using zero-order, first-order, parabolic-diffusion and modified Freundlich model. The correlation coefficient (R^2) and rate constants (k) are shown in Table 2.

Table 2. Correlation coefficient (R^2) and the rate constant (k) of the zero-order, first order, parabolic diffusion and the modified Freundlich kinetic model for both TiO₂ and TiO₂-AKC.

Kinetic models	TiO ₂ TiO ₂ -AKC			
	\mathbb{R}^2	k	\mathbb{R}^2	k
Zero order	0.023	-0.650	-1.45	0.668
(mgL ⁻¹ min ⁻¹)				
First order (min ⁻¹)	0.979	-0.015	0.936	-0.017
Parabolic diffusion	0.999	0.155	0.977	4.025
(mgkg ⁻¹ min ^{-1/2})				
Modified Freundlich	0.993	9.462	0.978	0.298
$(Lg^{-1}min^{-1})$				

Correlation coefficient (R^2) and the rate constant (k) of the zero-order, first order, parabolic diffusion and the modified Freundlich kinetic model for both TiO₂ and TiO₂-AKC.



Figure 6. Photocatalytic degradation kinetics plot for (A) zeroth order, (B) first order, (C) parabolic diffusion model and (D) modified Freundlich model of TiO₂.



Figure 7. Photocatalytic degradation kinetics plot for (A) zeroth order, (B) first order, (C) parabolic diffusion model and (D) modified Freundlich model of TiO₂-AKC.

The degradation kinetics model of TiO₂-AKC and TiO₂ could be described using first order, parabolic diffusion and Freundlich kinetic model. Though, parabolic diffusion best describes the kinetics of RhB in TiO₂ as shown in Table 2. This suggested that the process is controlled by heterogeneous diffusion similar to the report of Mukhlish *et al.* [16]. TiO₂-AKC composite fits into the modified Freundlich and parabolic diffusion model therefore; the process involves molecular ion exchange [16].

CONCLUSION

The prepared TiO₂-AKC composite was effective for photocatalytic degradation of Rhodamine B dye in aqueous solutions. The experimental results also revealed that TiO₂-AKC composite was more effective for degradation of Rhodamine B dye than sole TiO₂ using solar light. The experimental data analysed using zeroth and first order kinetics, parabolic diffusion, and modified Freundlich model conform into first order kinetics in TiO₂-AKC and sole TiO₂. The mechanism of the degradation process of RhB using TiO₂ fit best into parabolic diffusion model while TiO₂-AKC composite fit into the modified Freundlich and parabolic diffusion model. This finding also suggests that TiO₂-activated kaolinite composite is a good photocatalyst for the degradation of dyes.

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Samsudeen O. Azeez et al.

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