

Removal of cadmium (II) from aqueous solutions by two kinds of manganese coagulants

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Abstract

The study presents the removal of trace cadmium (II) in water with batch experiments by two kinds of manganese coagulants, HMO prepared from KMnO_4 and $\text{Na}_2\text{S}_2\text{O}_3$, and HFMO prepared from KMnO_4 and FeSO_4 . Various parameters such as turbidity, pH, initial concentration of Cd(II) and coagulation time were investigated. The results showed that the cadmium can be better removed using HMO than by HFMO. With pH = 6.72 and the Cd(II) initial concentration, 1mg/L, the dosage of HMO is 20mg/L, as the removal rate of cadmium is 85.04%. However, with pH = 8.03 and the Cd(II) initial concentration, 1mg/L, the dosage of HFMO is 20mg/L, as the removal rate of cadmium is 53.23%. The effectiveness of these two kinds of manganese coagulants for removal of cadmium and turbidity in natural water from Canal Jingmi in Beijing performed better when HMO is used than HFMO used. The paper indicated that HMO was an easily available and effective coagulant for the disposal of trace cadmium in water.

Keywords: Cadmium (II), HMO, HFMO, coagulation, binding site

1. Introduction

Contamination of the aqueous environments by heavy metals is a worldwide environmental problem due to their toxic effects and accumulation through the food chain. Cadmium is a heavy metal with the greatest potential hazard to humans and the environment due to its acute toxicity (Ozturk *et al.*, 2009). The health effects of Cd (II) on human include nausea, vomiting, diarrhea, muscle cramp, salivation, loss of calcium from bones, yellow coloration of teeth (cadmium ring formation), reduction of red blood cells, damage of bone marrow, hypertension, kidney failure following oral ingestion, lung irritation, chest pain, and loss of sense of smell after inhalation. Chronic cadmium poisoning produces proteinuria and affects the proximal tubules of kidney, causing formation of kidney stones (Naiya *et al.*, 2009). Drinking water regulations have continued to lower the maximum contaminant level (MCL) for cadmium based on more information about health effects. United States Environmental Protection Agency (USEPA) and the Ministry of Health of PR China have revised the (MCL) for Cd in drinking water from 10 $\mu\text{g/L}$ to 5 $\mu\text{g/L}$. The World Health Organization set the tolerance limits for cadmium concentration in drinking water as 3 $\mu\text{g/L}$.

The most commonly used methods for the removal of cadmium ions from aqueous solution include, chemical precipitation (Christensen *et al.*, 1982; Brantner *et al.*, 1984; Wu *et al.*, 2010), ion exchange (Bhattacharya *et al.*, 1984; Hruday, 1984), solvent extraction (Stenstrom, 1987; Casas *et al.*, 1986), dialysis/electro-dialysis (Marder *et al.*, 2003), reverse osmosis (Mazid *et al.*, 1984), supported liquid membrane (Tripathy *et al.*, 2002), and adsorption (Sharma *et al.*, 1998; Singh *et al.*, 1998). Among them, chemical precipitation, ion exchange and solvent extraction are not economically viable to remove cadmium to low concentrations from aqueous solutions. Adsorption is one of the most effective approaches to cadmium removal over a wider range of pH values and at lower concentrations. However, adsorption needs long contact time to achieve equilibrium between adsorbate and adsorbent, which restrains its practical application in water treatment (Gupta *et al.*, 2002; Mohapatra *et al.*, 2009; Das *et al.*, 2006). Therefore, the stiffening of regulations generates strong demands to improve methods for removing cadmium from the water and controlling water treatment residuals. Coagulation is one of the most important processes in water treatment plant; as well it is one of the most effective methods of the removal of cadmium (Ayoub *et al.*, 2001). Hydrous Manganese Dioxides (HMO) can exhibit

a low pH point of zero charge (pH_{PZC}) and can develop a large negative surface charge over the pH range of interest in most natural water (Tonkin *et al.*, 2004). Many researchers have studied the adsorption characteristics of cadmium by HMO. However, little attention was paid to the usage of hydrous manganese as a coagulant.

The present study deals with a series of batch coagulation experiments to investigate and explore the feasibility of hydrous manganese dioxide as a coagulant for removal of cadmium. The effect of turbidity, pH, and initial concentration of cadmium and coagulation time has been evaluated. Finally, the effectiveness of these two kinds of manganese coagulants was studied for removal of cadmium and turbidity in natural water from Canal Jingmi in Beijing.

2. Materials and methods

2.1 Raw water

All chemicals used in the experiment were of analytical grade. All solutions were prepared with deionized water. $CdCl_2 \cdot 5/2H_2O$ was dissolved in DI water to prepare primary stock solution with Cd(II) concentration of 1000mg/L every week. Hydrous manganese dioxide was employed as the coagulant in this study and was freshly prepared for each set of experiments. Background electrolyte solutions were prepared from the reagent-grade salts, NaCl and $NaHCO_3$ (Guan *et al.*, 2007). Turbidity of raw water was prepared by diatomaceous earth. Because the removal efficiency of cadmium by the manganese coagulant synthesized by potassium permanganate and ferrous sulfate was too low as the hydrogen ion generated, the raw water(2) spiked with 5mL, 0.1mol/L NaOH in the coagulation stage. The quality of raw water is listed in Table 1.

Table 1. The quality of raw water

NO.	pH	$Cd^{2+}/(mg/L)$	NaCl/(mol/L)	$NaHCO_3/(mol/L)$	Manganese coagulant
Raw water(1)	6.72	1	0.05	0.0005	$KMnO_4$ - $Na_2S_2O_3$
Raw water(2)	8.04	1	0.05	0.0005	$KMnO_4$ - $FeSO_4$

The natural raw water was collected from Canal Jingmi in summer, with turbidity of 0.6NTU, colority ranging from 1 to 2, COD_{Mn} of 1.85mg/L, and pH of 7.38. Because the concentration of cadmium in Canal Jingmi is low, the natural raw water is spiked with 1mL Cd^{2+} stock solution.

2.2 Materials

The manganese coagulants were prepared by the process $KMnO_4$ -Fe(\square) and $KMnO_4$ - $Na_2S_2O_3$, respectively.



HMO stands for the coagulant prepared by method (1). The primary ingredient of HMO is hydrous manganese dioxide, which accounts for 61.3% of the total amount of the precipitation (Yang *et al.*, 2008). HFMO stands for the coagulant prepared by method (2), which consists of hydrous manganese dioxide and ferric hydroxide. From the equation (2), the mass ratio of $MnO_2/Fe(OH)_3$ is 1/3.69. The dosage of HMO and HFMO is measured by the total amount of precipitation.

2.3 Methods

Jar tests were performed with a standard 1L jar testing device to simulate a conventional coagulation/flocculation process. All glassware was cleaned by soaking in 10% HNO_3 and rinsed three times with distilled water.

If not otherwise specified, the jar testing procedure was initiated with rapid mixing at 300rpm for 2min followed by 60rpm for 15min, consecutively. And then the samples were taken after 20min sedimentation. Rapid mixing was started as soon as the prepared manganese coagulants were added into synthetic water containing Cd(\square). The initial Cd(\square) concentration was 1mg/L and the coagulant was applied at 20mg/L. Each experiment was carried out open to the atmosphere in (at least) duplicate. All experiments were performed with a constant ionic strength of 0.05M NaCl. 0.0005M $NaHCO_3$ was added to provide the alkalinity. The pH values of the solutions were adjusted by adding 0.1 M sodium hydroxide and 0.1 M hydrochloric acid solutions. A high performance pH meter with a saturated KCl solution as electrolyte (Corning 350) was used to measure solution pH. Daily calibration with proper buffer solution (pH 4.00, 6.86 and 9.18) was performed to ensure its accuracy.

After reaction, the water was filtered immediately through 0.45 μ m pore size cellulose acetate membrane (MFS) and the filtrate was analyzed for remaining Cd concentration in the sample using atomic absorption spectrophotometry (AAS).

3. Result and Discussion

3.1 Removal efficiency by two manganese salts in synthetic water

The removal of Cd by two manganese coagulants over the dosage range of 2-50mg/L, was demonstrated in Figure 1. It was observed that the removal efficiency increased quickly as the increase of coagulant dose.

Adsorption is the main mechanism for the Cd(II) removal by manganese salts during coagulation. The pH_{pzc} of the HMO is found to be 1.8 (Tripathy *et al.*, 2006). So the HMO develops a large negative surface charge at pH 6.72. The Cd (II) was adsorbed onto the surface of HMO probably through electrostatic strength. The amount of the binding sites on the HMO surface increased with its dosage, consequently the removal efficiency increased quickly with the dosage of the HMO increasing. When the dosage of the HMO was more than 20mg/L, the remaining concentration of the residual cadmium ion was very low. And this restrained the adsorption of the cadmium. So the increase extent of the removal efficiency is small.

HFMO is constituted of ferric hydroxide and hydrous manganese dioxide. The pH_{pzc} of the HFMO is 4 (Guan *et al.*, 2009). It can be speculated that hydrous manganese dioxide plays a major role in the removal of cadmium. There are two aspects impact on the removal of cadmium by HFMO. On the one hand, removal efficiency increases with the HFMO dosage increasing, On the other hand, the removal of cadmium is highly pH-dependent.

The amount of hydrogen ions generated during the synthesis of HFMO increased with the HFMO dosage increasing. The removal efficiency of the cadmium declined with the decrease of the solution pH values. When the dosage of the HFMO was less than 28.14mg/L, the removal efficiency of the cadmium increased with the HFMO amount increasing. After the removal efficiency of the cadmium reached the plateau, it was mainly affected by the pH values of the solution and decreased with the decline of the pH values.

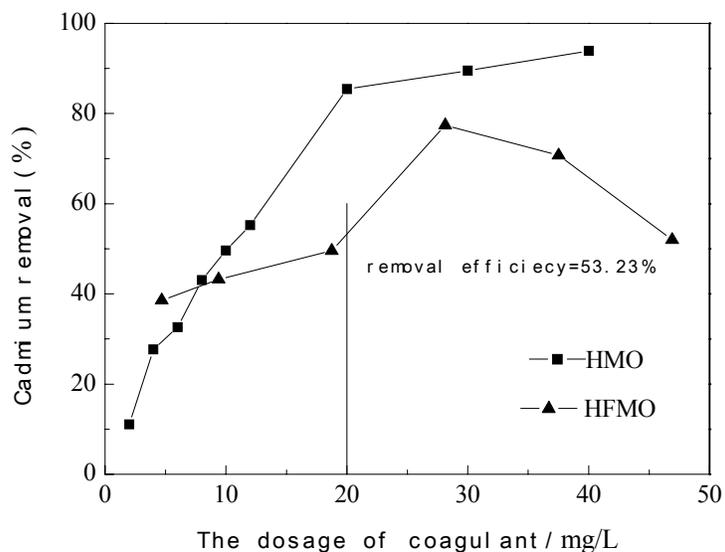


Figure 1. The removal efficiency of Cd(□) in synthetic water by HMO and HFMO

3.2 Effect of turbidity

Figure 2 shows the effect of the turbidity on removal efficiency of the Cd(II). It was observed that turbidity had little effect on the cadmium removal by HMO, Chen (2008) also reported that turbidity had no effect on the removal of cadmium by poly ferric sulfate. However, turbidity significantly promoted the cadmium removal by HFMO. When the turbidity was 3 NTU, the removal efficiency of cadmium is 97.57% increased by 20.16%. Over the turbidity range from 3NTU to 90NTU, the removal efficiency of the cadmium did not change greatly.

It is assumed that diatomaceous earth can absorb a certain amount of cadmium ions. But it takes up the same amount of binding sites. The adsorption capacity did not increase in the presence of the diatomaceous earth. As for HFMO, firstly the ferric hydroxide can absorb the diatomaceous earth; this led to the increase of the adsorption capacity of the HFMO. Secondly, diatomaceous earth adsorbed some hydrogen ions, which can be proved by the decrease of the pH value. This also increases the adsorption capacity of the HFMO. The increase of the turbidity did not increase the removal efficiency of the cadmium for the both manganese coagulants because the adsorption capacity for diatomaceous earth is also finite and the low concentration of residual cadmium restrains the adsorption of the cadmium.

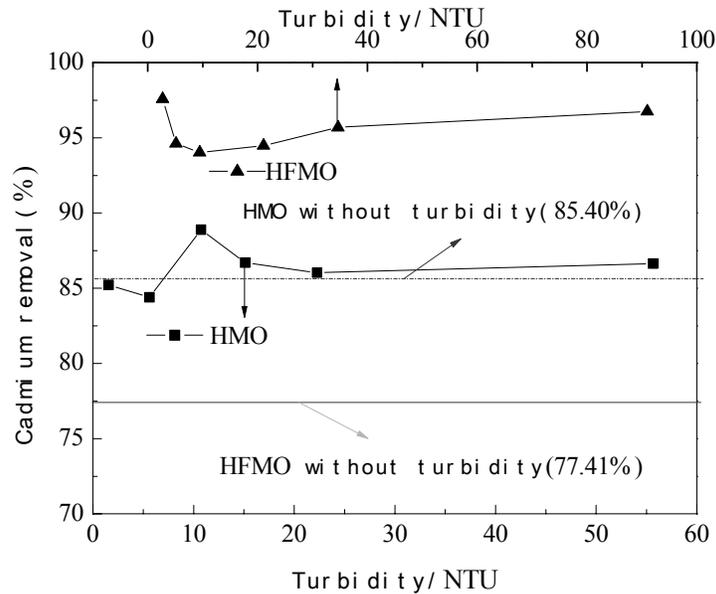


Figure 2. The effects of turbidity on the removal efficiency of Cd(II) in synthetic water by HMO and HFMO (The dosage of HMO and HFMO is 20mg/L, 28.14mg/L, separately)

3.3 Effect of pH

The pH of the solutions has an important variable governing metal removal. In general, removal of cations is favored at $\text{pH} > \text{pH}_{\text{pzc}}$. The effect of pH on the removal of Cd^{2+} by the two manganese coagulants was studied by varying pH of the solution over the range of 2~11 using different concentration. As demonstrated in Figure 3, the removals of cadmium by two manganese coagulants were highly pH-dependent. Cadmium removal at pH 2~4 could be negligible. Removal efficiency of the cadmium dramatically increased as the pH increased from 4 to 9. However, as the pH increased continuously the removal efficiency did not increase any more.

It is considered that alkalinity is beneficial to eliminate Cd^{2+} . When the $\text{pH} < 8$, cadmium mainly exists as Cd^{2+} in solution, the adsorption led to the increase of cadmium removal. The increased removal efficiency at $\text{pH} > 8$ may be a combination of both adsorption and precipitation on the surface of the coagulants. Firstly, the adsorption capacity increases with increase of the amount of surface charge. Secondly, at $\text{pH} > 8$, the fraction of $\text{Cd}(\text{OH})^+$, $\text{Cd}(\text{OH})_2$ in the solution increase with increase of pH. This also enhanced the cadmium removal.

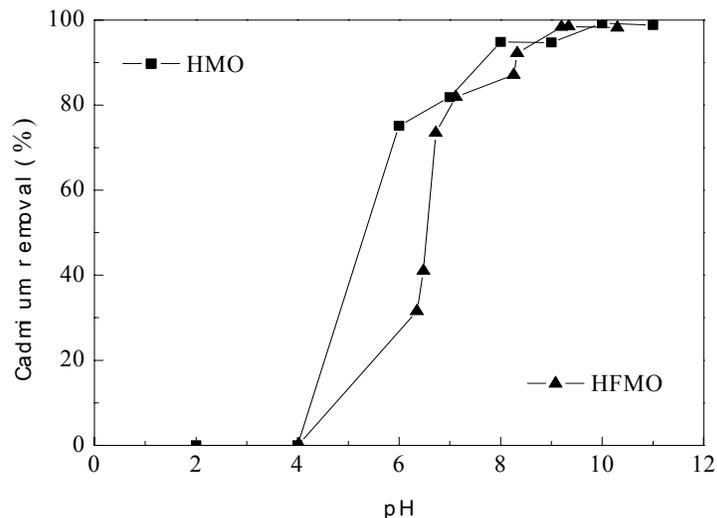


Figure 3. The effects of pH on the removal efficiency of Cd(II) in synthetic water by HMO and HFMO (The dosage of HMO and HFMO is 20mg/L, 28.14mg/L, separately)

3.4 Effect of initial Cd(II) concentration

Effect of initial Cd(II) concentration are shown in Figure 4. The cadmium removal increases with the increasing of the initial concentration, at the initial Cd(II) concentration less than 0.8 mg/L. With increase in initial Cd(II) concentrations, more Cd(II) ions are left in the solution due to the saturation of the binding sites. A maximum removal efficiency of cadmium is 94.08%, which was noted in Figure 4. As for HFMO, the removal of cadmium increased quickly over cadmium initial concentration 0.2-0.5mg/L. However, as the initial Cd(II) concentration increased continuously the removal efficiency did not increase any more. At the initial Cd(II) concentration over 1.13mg/L, the removal efficiency by HMO is larger than HFMO. Afterwards, the removal effect of cadmium by HFMO is better than HMO. This indicates that energetically less favorable sites become involved with increasing Cd(II) ion concentration in aqueous solution.

There are two kinds of binding sites, the lower and higher energy binding sites, on the surface of the coagulants. The binding sites with lower energy binding sites are available for Cd²⁺ when the initial Cd(II) concentration is high. The amount of highly energy binding sites of HMO is more than of HFMO. But the number of total binding sites of HMO is less than HFM (Tripathy *et al.*, 2006).

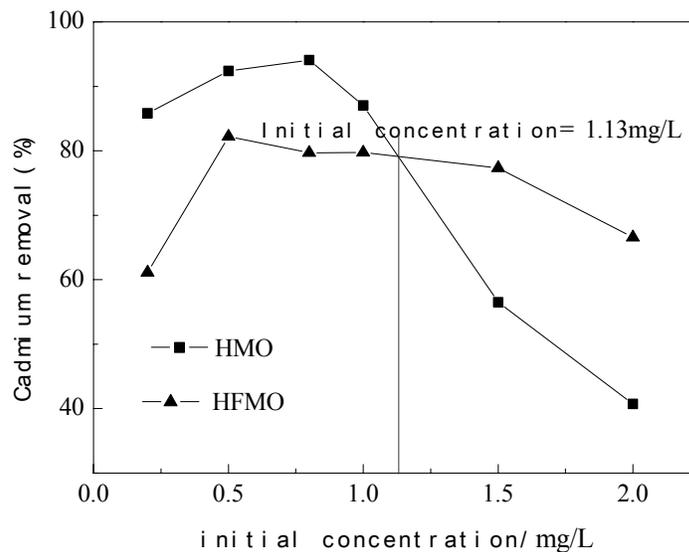


Figure 4. The effects of initial concentration of Cd²⁺ on the removal efficiency of Cd(II) in synthetic water by HMO and HFMO (The dosage of HMO and HFMO is 20mg/L, 28.14mg/L, separately)

3.5 Effect of coagulation time

In order to evaluate the effect of coagulation time on Cd(II) removal, coagulation experiments were conducted in the coagulation time from 2 min to 60min. As demonstrated in Figure 5, it is evident that Cd(II) can be removed very quickly at the rapid mixing stage. In the first 2 min, the removal of Cd(II) achieves 96.29% and 83.23% respectively by HMO and HFMO. Afterwards, the removal efficiency fluctuated within a small range.

3.6 Cadmium removal in the Canal Jingmi

Figure 6 shows the effect of coagulants dosage on the removal of cadmium by two manganese coagulants in natural water. It is observed that the removal efficiency variety in natural water is similar to the way in the synthetic water for HMO. When the dosage of the HMO is less than 8mg/L, the cadmium removal in natural water is better than in synthetic one because the alkalinity environment and the suspended substance enhanced the removal of the cadmium. But the cadmium removal in the natural water did not show any advantage after the dosage bigger than 8mg/L. This is due to the competing cations existed in the natural water inhibited the adsorption of the cadmium.

As far as HFMO, the removal efficiency in natural water is much higher than in synthetic water. This is also attributed to the alkalinity environment and the adsorption of the cadmium by suspended substance. The inserted figure shows that the turbidity after precipitation of the natural water treated by HFMO is very high and cannot satisfy the standards for drinking water quality. This restrains the application of the HFMO in water industry.

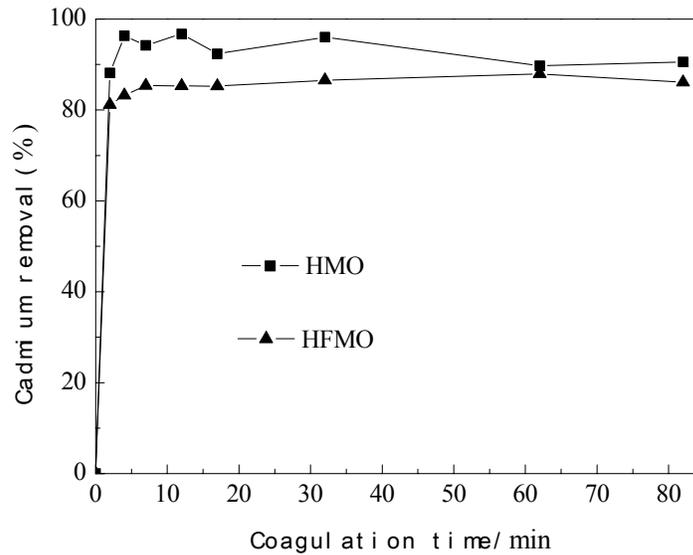


Figure 5. The effects of coagulation time on the removal efficiency of Cd(II) in synthetic water by HMO and HFMO (The dosage of HMO and HFMO is 20mg/L, 28.14mg/L, separately)

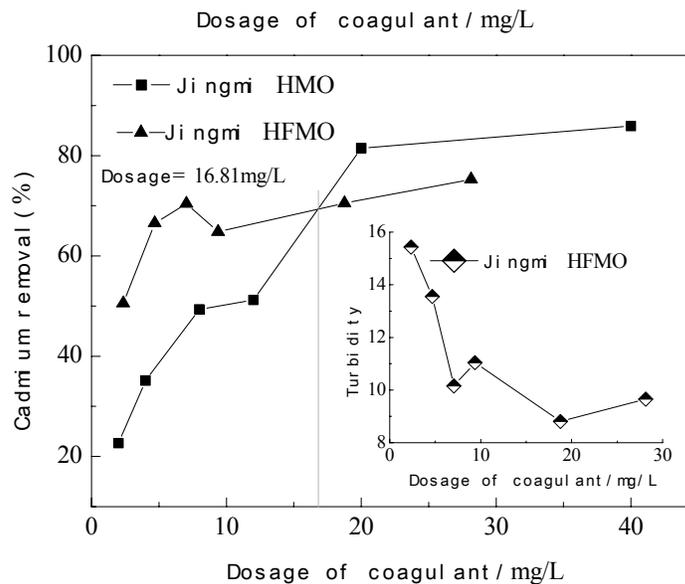


Figure 6. The removal efficiency of Cd(II) in raw water of Canal Jingmi of Beijing by HMO and HFMO (The inserted figure shows the turbidity after precipitation treated by HFMO)

4. Conclusions

This study examined Cd(II) removal by two different manganese coagulants. The obtained results can be summarized as follow:

- (1) HMO has a better cadmium removal efficiency and optimum Cd(II) removal by HMO was achieved at a HMO dosage of 20mg/L. As for HFMO, the removal efficiency of cadmium decreased when the dosage is bigger than 28.14mg/L.
- (2) Turbidity has little effect on the cadmium removal by HMO, but it significantly promotes the cadmium removal by HFMO. The change of the turbidity has slightly effect on the cadmium removal for both manganese salt coagulants.
- (3) The cadmium removal is highly pH-dependent. The removal efficiency of the cadmium is bad under the acidic environment. However, the removal efficiency increases quickly with increasing of the pH.
- (4) The initial concentration of the cadmium significantly affects the cadmium removal by HMO and optimum cadmium (II)

removal was achieved at a initial concentration of 1mg/L. However, the initial concentration has slightly effect on the cadmium removal by HFMO.

- (5) Cadmium (II) can be removed very fast and most of it is removed at the rapid mixing stage. When the coagulation time is just 2 min, the percentage removal of Cd(II) by HMO and HFMO reached 96.29% and 83.23%, respectively.
- (6) The removal efficiency change in natural water is similar to the way in the synthetic water for HMO. As far as HFMO, the removal efficiency in natural water is much higher than in synthetic water. But the turbidity after precipitation of the natural water treated by HFMO is very high and cannot satisfy the standards for drinking water quality.

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Received April 2010

Accepted September 2010

Final acceptance in revised form September 2010