This study was undertaken to investigate the effects of Bonny light crude oil contamination of sandy loam soil on aspects of microbial metabolism and physicochemical properties of the soil. Bonny light crude oil (specific gravity = 0.81) was used at eight different levels (0.5%, 1.0%, 2.0%, 2.5%, 5.0%, 10.0%, 15.0% or 20.0% v/w of soil) for the controlled pollution of pristine soil samples, each weighing 1 kg. The experiment lasted for eight weeks. Results of the effects of crude oil on the physicochemical properties of the soil showed that high levels of the oil significantly ($p< 0.05$) increased soil organic matter but had no significant effect on the pH and moisture content. With the exception of organic carbon, the levels of bioavailable nitrogen, sodium, potassium, calcium, magnesium, sulphur and phosphorus in the test samples with higher levels of crude oil (5.0%, 10.0%, 15.0% and 20.0%) were significantly reduced when compared to their levels in the controls. Similarly, higher levels of the oil significantly ($p<0.05$) reduced soil microbial phospholipid synthesis and CO$_2$ emission. Correlation analysis using the Pearson's correlation model showed a positive correlation between soil CO$_2$ and phospholipid ($r = 0.74$).

**Keywords:** Contamination, Crude oil, Microbial respiration, Physicochemical properties.

**INTRODUCTION**

Crude oil pollution of the environment is a global problem which has meted out untold havoc on the ecosystem. Even though crude petroleum since its discovery as an energy source has greatly improved man’s living conditions, the ever-increasing damages it wreaks on the ecosystem underscore its role as a potent environmental pollutant. It has been shown to negatively affect both the biotic and abiotic components of the ecosystem (McOrist and Lenghaus, 1992; Chaineau *et al*., 2000; Adoki and Orugbani, 2007; Mariana *et al*., 2010; Eze *et al*., 2013; Sulaiman *et al*., 2015; Ikuesan, 2017; Odo *et al*., 2019). The toxicity of crude and refined oil products to living organisms is an issue that needs urgent attention, especially in oil-rich nations of the world. In areas of heavy oil exploration and exploitation, ecologists have noted a drastic decline in the population of many plant and animal species as well as a sharp depletion in the fertility of agricultural lands (McOrist and Lenghaus 1992).

In Nigeria for instance, oil spill is a major environmental problem. According to information from the Department of Surveying and Geoinformatics, University of Lagos, between 1976 and 1996, Nigeria recorded a total of 4,835 oil spill incidents which resulted in the loss of 1,896,960 barrels of oil to the environment. In 1998, still in Nigeria, 40,000 barrels of oil from Mobil platform off the Akwa Ibom coast were spilled into the environment causing serious damage to the coastal environment. Oil spillage along the Nigerian coasts has severely affected the flora, fauna, drinking water as well as human lives and property in those regions. Because of the perennial nature of crude oil pollution, painstaking research on the mechanisms by which it negatively affects living and non-living matter is necessary to help design remediation techniques for contaminated environments. Many researchers have delved into the effects of crude oil pollution on living organisms but to date, there is paucity of research information on the effects of this natural resource on arable lands. To this end this study was undertaken to investigate some of the ways by which crude oil negatively impacts on the soil as well as on microbial respiration.
MATERIALS AND METHODS
The major materials used in the study were:

1) Bonny light crude oil which was obtained from the Nigerian National Petroleum Corporation (NNPC) Port Harcourt Refinery, Alesa-Eleme, Rivers State, Nigeria.

2) Pristine sandy loam soil collected from Botanical Garden, University of Nigeria, Nsukka.

Experimental Design
The study was grouped into two phases. The first phase investigated the effects of crude oil on microbial metabolism which consisted of its effects on soil microbial respiration and phospholipids synthesis. The second phase was about the effects of crude oil on the physicochemical properties of the soil.

Bonny light crude oil (specific gravity = 0.81) was used at eight different levels (0.5%, 1.0%, 2.0%, 2.5%, 5.0%, 10.0%, 15.0% or 20.0% v/w of soil) for the controlled contamination of pristine soil samples, each weighing 1 kg. The control soil samples were not contaminated. Each experiment lasted for eight weeks.

Determination of the Effects of Crude Oil on Soil Microbial Respiration
This was done by following the method of Isermeyer (1952) which quantified the level of CO₂ evolved from the soil. Fifty grams (50 g) of each soil sample was weighed in duplicate into beakers placed inside jars with air-tight covers. Twenty-five millilitres (25 ml) of 0.05 M NaOH was pipetted into each jar and immediately the jars were made airtight with rubber rings. Three jars containing 0.05 M NaOH but without soil were used as controls for both contaminated and uncontaminated soil samples. Each weighing 1 kg. The control soil samples were not contaminated. Each experiment lasted for eight weeks.

Estimation of CO₂
At the end of the incubation, the jars were opened and the beakers taken out. The external surface of each beaker was washed with CO₂-free water (prepared by cooling boiled distilled water in a container with CO₂ absorption tubes) to wash the NaOH solution completely into the jar. Thereafter, 5 ml of 0.5 M barium chloride solution was added to each jar together with some drops of phenolphthalein indicator. A 0.05 M HCl was added drop wise with continuous stirring until the colour changed from red to colourless. Calculation of results:

The rate of the respiration was calculated by the following relationship:

\[
\frac{\text{CO}_2 (\text{mg})}{\text{SWt}} = \frac{(V_0 - V)}{\text{DWT}} \times 1.1
\]

Where SWt is the amount of soil dry weight in grams, T is the incubation time in hours, V₀ is the volume of HCl used for blank titration (average value) in milliliters, V is the volume of HCl used for the soil sample (average value), DWT is the dry weight of 1 g moist soil and 1.1 is the conversion factor (1 ml 0.05 M NaOH equals 1.1 mg CO₂).

Determination of Microbial Phospholipids
This was done by the method of Frostegard et al. (1991). Glassware was washed in methanol, 15% HNO₃ and rinsed twice in tap water and three times in deionized water. Extraction of phospholipids was achieved using a chloroform: citrate buffer (1:2:0.8, v/v/v) as extractant. The citrate buffer contained 0.015 M citric acid and 0.15 M trisodium citrate (5.9:4.1 v/v) producing a pH of 4.0. One gram of wet weight of soil was placed in a McCartney bottle with 11.65 ml of extractant. The ratio of chloroform to soil was 3:1; within the range of 2 – 6: 1 used by Frostegard et al. (1991). After 2 hours, 3.1 ml of citrate buffer and 3.1 ml of chloroform were added and the solution periodically shaken throughout a further hour to assist extraction. Solutions were then left overnight to separate into two phases. Part of the lower chloroform phase (6.4 ml final volume) was withdrawn with a pipette, using care to avoid the inclusion of any soil particles and placed in a small vial. A 0.1 – 1.0 ml aliquot was then transferred to a 5 ml bijou bottle and dried with nitrogen gas.

Digestion and phospholipids analysis were performed as described by Findlay et al. (1989). Lipid extracts and glycerol phosphate standards were digested with 1.8 ml acidified potassium persulphate (K₂S₈O₉) solution (5 g to 100 ml of
0.35 N \( \text{H}_2\text{SO}_4 \) for 24 h at 95 °C. While still hot, 0.4 ml of ammonium molybdate solution (2.5% \((\text{NH}_4)\text{MoO}_4\) in 5.72 N \( \text{H}_2\text{SO}_4 \)) was added and left for 10 min before the addition of 1.8 ml of malachite green solution. Absorbance at 610 nm was read after 30 min. Distilled water was used as blank.

**Determination of Total Organic Carbon and % Organic Matter in Soil**

The method of Walkley–Black (1934) was followed. One gram of each soil sample was weighed into a 250 ml conical flask. A 10.0 ml 1 N \( \text{K}_2\text{Cr}_2\text{O}_7 \) solution was pipetted accurately into each flask and swirled gently to disperse the soil. Thereafter, 20 ml of 0.05 M \( \text{H}_2\text{SO}_4 \), was rapidly added by means of a pipette. The flask was immediately swirled until soil and reagents were mixed and left to stand on white tile for about 30 minutes. This was followed by addition of 100 ml of distilled water. Finally three drops of orthophenanthroline indicator (made by dissolving 14.8 g of 0-phenanthroline monohydrate and 6.9 g of \( \text{FeSO}_4\cdot7\text{H}_2\text{O} \) in 1 litre of distilled water) was added. This was titrated with acidified 0.5 N ferrous sulphate solution (14 g/L of water containing 15.0 ml 0.05 M \( \text{H}_2\text{SO}_4 \)). The ferrous sulphate was added drop by drop until the colour changed from blue to red (maroon colour) in reflected light against the white background.

The results were calculated according to the following formula:

\[
\% \text{Organic carbon in soil} = \frac{(\text{MeKCrO}_7 – \text{MeFeSO}_4)}{\text{g of air – dry soil}} \times 0.003 \times 100 \times F \\
\text{Correction factor, } F = 1.33 \\
\text{Me} = \text{Normality of solution x ml of solution used.}
\]

% organic matter was determined by multiplying percentage organic carbon (\(\%\text{C}\)) by 1.724, that is, % organic matter = \(\%\text{C} \times 1.724\).

**Total Nitrogen Determination**

The method of Keeney and Nelson (1982) was followed. Five grams of air-dried and sieved (2 mm sieve) soil sample was fed into a 500 ml Kjeldahl flask. A catalyst mixture with the following composition was prepared: 0.20 g selenium powder, 5.0 g lithium sulphate and 100 ml of 30% hydrogen peroxide. Two hundred millilitres (200 ml) of conc. sulphuric acid was mixed with the catalyst and 100 ml of this digestion mixture was added slowly to the sample in the Kjeldahl flask. The entire mixture was heated in a fume-cupboard until the digest was clear. This lasted for 2 h. The digest was allowed to cool and subsequently 50 ml was transferred into a 1000 ml Kjeldahl flask containing 400 ml distilled water. It was mixed thoroughly and allowed to settle until a clear solution was formed. One hundred milliliters (100 ml) of 45% sodium hydroxide was added to the solution in the flask to make it alkaline and the flask immediately connected to the distillation apparatus. The ammonia was collected over 50 ml of 2.5% boric acid solution contained in a 250 ml conical flask with 5 drops of mixed indicator (bromocresol green and methyl red mixed in ethanol). Two hundred milliliters (200 ml) of the distillate was collected and titrated with standard 0.05 N HCl.

\[
\% \text{Nitrogen was calculated with the formula:} \\
\frac{T \times N \times 14.01 \times 100}{1000 \times W_1}
\]

Where:

\(T\) = Sample titre
\(N\) = Acid normality
\(W_1\) = Weight of soil sample used

**Determination of \(\text{Ca}^{2+}\), \(\text{Na}^+\), \(\text{Mg}^{2+}\), and \(\text{K}^+\)**

The method of Black (1965) and AOAC (1970) was adopted in this analysis. Five grams of each soil sample was weighed into Whatman No. 1 filter paper fitted into a funnel on a leaching stand with 100 ml volumetric flask to collect the leachate. The sample was leached with 1.00 N ammonium acetate solution and 100 ml volume was collected.

For the determination of \(\text{Ca}^{2+}\) and \(\text{Mg}^{2+}\), ten milliliters (10 ml) of the leachate was dispensed into a 100 ml conical flask. This was followed by the addition of 10 ml ammonium chloride solution to the flask.

The resulting solution was mixed well and 0.01g of Eriochrome Black-T indicator was incorporated and titrated immediately with 1 N
EDTA in the burette.

Calculation:
\[ T \times N \times \frac{V_{ol}}{A_{liq.c}} \times \frac{100}{100 \text{ g of soil} \times W_{t}} = \text{Meq. Ca}^{2+} \text{Mg}^{2+} \]

Where
- \( T \) = Sample titre
- \( V_{ol} \) = Volume of leachate collected originally
- \( A_{liq.c} \) = Aliquot of leachate titrated
- \( W_{t} \) = Weight of soil sample used
- Meq. = Milliequivalent

For \( \text{Ca}^{2+} \) determination, ten milliliters of the leachate was pipetted into another 100 ml conical flask. Thereafter 20 ml of 20% potassium hydroxide was added and 0.01g calcien indicator incorporated; the resulting solution was titrated with 0.01 N EDTA in a burette.

Calculation:
\[ \text{Meq. Ca}^{2+} / 100 \text{ g of soil} = T \times N \times \frac{100}{W_{t} \times A_{liq.c}} \times V_{ol} \]

Where:
- \( T \) = Sample titre
- \( N \) = Normality of EDTA
- \( W_{t} \) = Weight of soil sample used

Meq. \( \text{Mg}^{2+} \) was obtained by subtracting meq. \( \text{Ca}^{2+} \) from meq. \( \text{Ca}^{2+} + \text{meq. Mg}^{2+} \), that is, meq. \( \text{Ca}^{2+} + \text{Mg}^{2+} - \text{meq. Ca}^{2+} \),

Sodium (\( \text{Na}^{+} \)) and potassium (\( \text{K}^{+} \)) were determined colorimetrically using a flame photometer (Jenway PEP7 Flame Photometer, Germany) from the remaining leachate.

**Sulphur Determination**

For the determination of sulphur, the method of Tabatabai and Bremner (1970) was adopted. Soil sample was finely ground and 0.1 g was weighed into a ceramic crucible. Subsequently 3 ml sodium hypobromite solution (NaOBr) was added and allowed to stand for 5 min. The digestion mixture was placed on a laboratory hot plate and evaporated to dryness at 250 °C. After cooling to room temperature, 3 ml distilled water was added and the digest heated again for 30 s. This was followed by agitation to dissolve the residue and then cooling to ambient temperature. It was then transferred quantitatively to a 10 ml volumetric flask and more water was added to make the solution to volume in the volumetric flask. Particles that could interfere with the analysis were filtered off before the absorbance was read at 420 nm in a colorimeter (Jenway 6051, Germany).

**Determination of Phosphorus in Soil Sample**

The method of Black (1965) was followed. Two grams of air-dried and sieved soil was weighed into a test tube and to the test tube was added 20 ml of 0.03 N ammonium fluoride (NH₄F) in 0.1 N HCl and the tube stoppered. The tube was agitated for 1 min, allowed to settle and subsequently filtered using a filter paper. One milliliter (1 ml) of the filtrate was pipetted into another test tube. To this test tube was added 7.0 ml of distilled water, 1.0 ml of 2.5% ammonium molybdate and 1.0 ml of 2% ascorbic acid, bringing the total volume to 10 ml. The solution was mixed very well and placed in a beaker of boiling water for 3 min. The optical density was read at 620 nm using a colorimeter (Jenway 6051 Colorimeter, Germany).

**Soil Moisture Content Determination**

Two porcelain basins were weighed and the weights recorded. Twenty (20 g) grams of each of the wet soil samples (both the control and polluted samples) was weighed in duplicate into each of the basins. The samples were dried in an oven at 105 °C for 24 h and later cooled in a desiccator. The dry soil samples were re-weighed and the weights obtained by subtracting the weight of the empty basin from the combined weight of the basin and the dry soil.

\[ \text{The % soil moisture content} = \frac{\text{Moist weight} - \text{Dry weight}}{\text{Moist weight}} \times 100 \]

**Soil pH in Distilled Water**

The method of Black (1965) was followed. A 20.0 g sample of the soil was put into a 50 ml beaker containing 20 ml of distilled water. The mixture was allowed to stand for 30 min, stirring occasionally with a glass rod. The electrodes of the pH meter were inserted into the partly settled suspension and the pH measured.
Determination of the Levels of Transition Elements in the Crude Oil

This was done following the methods of Black (1965). One millilitre (1 ml) of the crude oil was pipetted into a 250 ml conical flask and 40 ml of 1:1 HNO₃/perchloric acid was added and allowed to stand for 24 hr. The digestion mixture was placed on a heater inside a fume cupboard and heated to eliminate the carbon content until white fumes appeared indicating complete digestion. It was heated further to a near dryness and cooled. Distilled water was added and 100 ml of the digest filtered into a volumetric flask. The filtrate was used for the determination of the levels of iron, cobalt, nickel and copper.

(a) Iron Determination (Orthophenanthroline Method)

Ten milliliters (10 ml) of the crude oil extract was pipetted into a 50 ml volumetric flask. To the extract was added 2 ml of 10% hydroxylamine hydrochloride solution, 20 ml of 10% sodium citrate and 2 ml of 0.25% orthophenanthroline. The solution was made up to 50 ml volume with distilled water and allowed to stay for 24 h for colour development.

The absorbance was read at 510 nm and the quantity of iron in mg/L of the crude oil calculated as:

\[
\text{Quantity in mg/L} = \frac{\text{Absorbance} \times \text{Total extract}}{\text{Aliquot taken} \times \text{Weight of sample}} \times \text{Gf} \times \frac{1}{\text{Gf}}
\]

Where:

\[ \text{Gf} = \text{slope of standard curve prepared by plotting the absorbance of pure iron metal against their concentrations} \]

(b) Nickel Determination (Back-titration with ZnSO₄)

Ten milliliters of the extract was introduced into a 250 ml conical flask containing 50 ml of distilled water. Thereafter, 10 ml of ammonia buffer solution, 1 ml of 0.05% orthophenanthroline, 5 ml of 0.01 N EDTA and a pinch of Eriochrome Black-T indicator were added. This was back–titrated with 0.01 N ZnSO₄ to a reddish end point.

Calculation:

\[
\text{Quantity of nickel in the crude oil (mg/L)} = \frac{\text{Titre} \times \text{normality of metal (ZnSO₄)} \times \text{Equivalent weight of metal} \times \text{Total extract}}{\text{Aliquot taken} \times \text{Weight of sample}}
\]

Where:

\[ \text{Equivalent weight of metal} = \frac{\text{Molar mass of metal}}{\text{Valency of metal}} \]

(c) Determination of Cobalt

Ten milliliters (10 ml) of the crude oil extract was added to a 250 ml conical flask containing 50 ml of distilled water. Thereafter, 10 ml of ammonia buffer solution, 1 ml of 0.05% orthophenanthroline, 5 ml of 0.01 N EDTA and a pinch of Eriochrome Black–T indicator were added. This was back – titrated with 0.01 N ZnSO₄ to a reddish end point.

Calculation:

\[
\text{Quantity of cobalt in the crude oil (mg/L)} = \frac{\text{Titre} \times \text{normality of metal} \times \text{Equivalent weight of metal} \times \text{Total extract}}{\text{Aliquot taken} \times \text{Weight of sample}}
\]

Where:

\[ \text{Equivalent weight of metal} = \frac{\text{Molar mass of metal}}{\text{Valency of metal}} \]

(d) Copper Determination

Ten milliliters (10 ml) of the crude oil extract was pipetted into a test tube. To the test tube were added 5 ml of 2 N NH₄OH solution and 5 ml of 0.05 N potassium ferrocyanide solution. After mixing, the test tube was allowed to stand for 2 h, and then the absorbance was read at 400 nm.

Calculation

\[
\text{Total copper in the crude oil (mg/L)} = \]

The excess EDTA was titrated with N ZnSO₄ solution until the colour changed from blue to wine red end point.
Absorbance x slope of standard curve x Total extract
Aliquot taken x Weight of sample

Statistical Analysis
Data analysis was carried out using a one-way analysis of variance (ANOVA) and the difference was done by comparing tests with P<0.05.

RESULTS
Effects of Crude Oil on Soil Respiration (Microbial Respiration) and Phospholipid Levels
Results of these assays are presented in figures 1 and 2. Soil respiration was quantified in terms of the level of CO\textsubscript{2} evolved from the soil over a specified period of time. High levels of crude oil significantly (P<0.05) reduced the levels of CO\textsubscript{2} evolution (Figure 1) and phospholipids (Figure 2) in the soil samples. Exposure time of the crude oil to microorganisms positively affected the levels of CO\textsubscript{2} and phospholipids in the soil. For instance, the levels of both substances after eight weeks of incubation were much higher than their levels after four weeks (Figures 1 and 2). The level of CO\textsubscript{2} in the control sample after four weeks of incubation was higher than the levels in the polluted (test) samples. However after eight weeks of incubation the level of CO\textsubscript{2} in the control became lower than its level in each of the test samples (Figure 1). Maximum CO\textsubscript{2} levels in the test samples occurred at 0.5% and 1% crude oil levels in both the four-week and eight-week incubated samples. This trend also occurred with the phospholipid levels - maximum levels were also obtained at 0.5% and 1% crude oil concentrations. Correlation analysis using the Pearson’s correlation model (Figure 3) showed a positive correlation between soil CO\textsubscript{2} and phospholipids (correlation coefficient = 0.74).

Figure 1: Effects of Crude Oil on Soil Respiration
Figure 2: Effects of Varying Levels of Crude Oil on Soil Microbial Phospholipids (nm/g of soil)

Figure 3: Relationship between Microbial Phospholipid and CO$_2$ Evolution from Crude Oil contaminated Soil Samples.

Effects of Crude Oil on Soil pH, Moisture and Organic Matter Levels
These are depicted in figure 4. With the exception of pH which was almost stable in all the levels of crude oil contamination, soil moisture and organic matter increased with increase in crude oil level.
**Effect on Soil Macronutrients—Carbon, Nitrogen, Calcium, Sodium, Potassium, Phosphorus, Magnesium and Sulphur**

Soil organic carbon (Figure 5) increased with increase in crude oil level but the reverse was the case with nitrogen and other macronutrients which decreased as crude oil dose increased (Figures 6 and 7).

---

**Figure 4:** Effects of Crude Oil on pH, Moisture and Organic Matter Levels of the Soil

**Figure 5:** Effect of Crude Oil on the Percentage Composition of Soil Carbon and Nitrogen
Quantitative Composition of Some Transition Elements in the Crude Oil Sample
This is presented in table 1. The crude oil contained significant (P<0.05) levels of iron, copper, cobalt and nickel.

Table 1: Quantitative Composition of Some Transition Elements in the Crude Oil Sample

<table>
<thead>
<tr>
<th>Elements</th>
<th>Fe</th>
<th>Cu</th>
<th>Co</th>
<th>Ni</th>
</tr>
</thead>
<tbody>
<tr>
<td>Composition (mg/L of crude)</td>
<td>2.995±0.01</td>
<td>5.194±0.11</td>
<td>1.179±0.02</td>
<td>1.467±0.13</td>
</tr>
</tbody>
</table>
DISCUSSION
Crude oil had significant effects on some physicochemical properties of the soil. Organic matter level, for instance, increased significantly (P<0.05) with increase in crude oil level, even though this increase was caused by the introduction of organic carbon into the soil by the crude oil. Increase in organic matter under normal circumstances is expected to improve soil fertility but in this case because the increase was crude oil-associated it rather decreased soil fertility. This is because the oil seals up soil pore spaces preventing water and air movement into the soil, lowers biological activity and disrupts biochemical conversions necessary to make mineral nutrients available to plants. Soil pH and moisture were not significantly affected by the oil.

Crude oil had a dose-dependent inhibitory effect on the levels of macroelements in the soil, with the exception of organic carbon which increased progressively with increase in crude oil level. There were significant reductions (P<0.05) in the levels of soil N, Na, K, Ca, and S as crude oil level increased, when compared to their levels in the uncontaminated control (Figures 5-7). It was only in magnesium and phosphorus that the effect was not pronounced. This may be because crude oil binds most of these mineral nutrients making them unavailable. A likely mechanism through which crude oil does this is through the transition element constituents. The crude oil is rich in transition elements (Table 1) and these elements are characterized by the possession of partially filled $d$ or $f$ orbitals in any common oxidation state. Through these free orbitals, they form coordinate bonds and coordinate complexes with ligands (eg. $\text{NH}_3$, $\text{H}_2\text{O}$, $\text{NO}_3^-$ etc) or other elements (eg K, Ca, Mg etc) (Coxon et al., 1980). When such complexes are formed, the bound ions will lose their ionic properties and will not be detectable in solution. Okolo et al. (2005) also reported that oil pollution reduced the levels of soil nitrate and phosphorus but they did not investigate the effects on other macroelements. Wyszkowski et al. (2001) had earlier reported an increase in the levels of nitrogen, phosphorus and potassium in crude oil-polluted soil but in their case the soil was amended with NPK compound fertilizer.

The oil reduced the rate of soil respiration and the level of microbial phospholipids in a dose-dependent manner (Figures 1 and 2). Phospholipids are a component of all cell membranes and they remain in constant proportion to the bacterial biomass and rapidly disappear after cell death (Peterson et al. 1991). According to Frostegard et al. (1991), phospholipid levels are used to measure total microbial biomass, activity and metabolic status. Correlation analysis using the Pearson’s correlation model (Figure 3) showed a positive correlation between soil CO$_2$ and phospholipid (correlation coefficient = 0.74).

CONCLUSION
This study has revealed some of the deleterious effects of crude oil pollution on soil microbial respiration and physicochemical properties. Further research into more of the mechanisms by which this natural resource impoverishes the soil will help environmental managers to design strategies for the remediation of crude oil contaminated soils.

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