

RESEARCH PAPER
**COLUMN STUDIES ON THE REMOVAL OF CHROMIUM
FROM WASTE WATER BY MANGO SEED SHELL
ACTIVATED CARBON**

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ABSTRACT

The effects of time and carbon height on the removal of chromium from wastewater were investigated in fixed down flow adsorption columns containing mango seed shell activated carbon (MSSAC) for the purpose of converting the waste to wealth. The Hutchin's bed depth service time (BDST) model was used to study the columns performance at 10% and 90% breakthrough concentrations. The BDST model constants were determined. The BDST equation obtained at flow rate of 1.6l/hr and influent chromium concentration of 3.151mg/L was used to predict the adsorbent performance at other flow rates and initial chromium concentrations. Results obtained revealed that there was a gradual decrease in the percentage of chromium adsorbed with time; the percentage adsorbed became zero at 72.33, 467.5 and 572 hours respectively for 3cm, 3.5cm and 4cm carbon heights. The breakthrough curves for chromium adsorption deviated from the characteristic S-curve for most dynamic studies because 100% removal of chromium was not attained even at the commencement of the column operation because of the high initial concentration of the adsorbate adopted in this study. It was observed that for a given breakthrough concentration, the service time decreased with increase in both flow rate and initial chromium concentration. Chi-square test performed on the developed BDST models revealed the reliability of the models for the prediction of the columns performance particularly at lower breakthrough concentrations. It was therefore recommended that MSSAC can be used for the removal of chromium from wastewater in adsorption columns; however there is the need for pilot studies on live wastewater to investigate the effect of interference.

Keywords: Column Adsorption, Activated carbon, mango seed shells, wastewater, pilot studies

INTRODUCTION

The sorption capacity of a given sorbent for pollutants is usually determined by carrying out an adsorption isotherm. However, isotherms cannot give accurate scale-up data in fixed or fluidised bed processes, because adsorption in a

flow column is not at equilibrium. The optimum operating capacity and contact time must be determined in vivo to decide upon the best column dimensions and the number of units needed for continuous treatment (McKay and Bino, 1990).

Therefore, for an optimal design of an industrial adsorption process, it is important to have accurate modelling and simulation of the dynamic behaviour of the fixed bed system. Several models have been proposed to predict results for a variety of adsorption systems. These models based on fundamental mass transport mechanisms including external film; pore and surface diffusion require the solution of a number of non-linear partial differential equations which include physical and kinetic parameters. These equations can be solved only by numerical methods that are time consuming and tedious.

By using simplified models, pilot plant testing could be largely used for verification rather than information gathering, saving time and money. The bed depth service time (BDST) model (Hutchins, 1973), mass transfer zone model (Michaels, 1952) and Thomas model (1948) among others offer simplified approach and rapid prediction of adsorber performance.

THEORETICAL-BDST MODEL

Bohart and Adams (1920) developed a relationship based on surface-reaction-rate theory, which can be used to predict the performance of continuous carbon columns. Hutchins (1973) proposed a simple approach based on Bohart and Adams (1920) model for fixed bed adsorbents to correlate the service time with the process variables. The linearized form of this relationship is given thus (Hutchins, 1973):

$$t = \frac{N_o H}{C_o q} - \frac{1}{k C_o} \ln \left(\frac{C_o}{C_b} - 1 \right) \quad (1)$$

Where:

t is service time (h) to breakthrough; N_o is adsorption capacity (mg solute/l adsorbent); C_o is initial solute concentration (mg/L); q is linear flow rate (m/h); H is depth of adsorbent bed (m); k is rate constant of adsorption (l/mg.h) and C_b is outlet concentration at desired breakthrough level (mg/L).

By plotting t against H from experimental data, N_o can be evaluated from the slope of the

graph and k from the intercept at t = 0. The bed depth, which theoretically is just sufficient to prevent concentration in excess of C_b at t=0, is defined as the critical bed depth (H_o) of the column and is determined from Equation (1) when t = 0 thus:

$$H_o = \frac{q}{k N_o} \ln \left(\frac{C_o}{C_b} - 1 \right) \quad (2)$$

At least nine individual column tests must be conducted to collect the laboratory data required for the Bohart–Adams approach, an expensive and time-consuming task. While with Hutchins modification of Bohart–Adams equation, only three fixed bed tests are required to collect the necessary data instead of nine columns. Another advantage of the Hutchins approach is that, if a value of the slope for one flow rate is obtained, values for other flow rates will be computed by multiplying the original slope by the ratio of the original and new flow rates. The intercept, b, value change is insignificant with respect to changing flow rates. Adjustment for changing initial concentration can also be made as follows (Eckenfelder, 1989):

$$a_2 = a_1 \frac{C_1}{C_2} \quad (3)$$

$$b_2 = b_1 \frac{C_1 \left[\ln \left(\frac{C_2}{C_f} - 1 \right) \right]}{C_1 \left[\ln \left(\frac{C_1}{C_b} - 1 \right) \right]} \quad (4)$$

Where:

C_f and C_b are the effluent concentrations with respect to C_2 and C_1 respectively. But for a fixed breakthrough effluent concentration, $C_f = C_b$. Hence Equation (4) can be used to evaluate b_2 .

A survey of literature revealed that Zhao and Duncan (1998) produced activated carbon from *Azolla filiculoides* and used it in fixed-bed adsorption columns for quantitative determination

of the characteristic process parameters which could then be used for performance comparison and process design for zinc removal. Mazumder *et al.*, (2011) studied the removal of colour, copper and chromium from electroplating wastewater in adsorption columns using materials such as wood charcoal, coconut shell carbon and activated alumina. They reported that activated alumina removed the pollutants better than the other carbons.

Lim *et al.* (2008) studied the adsorption of lead (Pb^{2+}), copper (Cu^{2+}) and zinc (Zn^{2+}) on sawdust in fixed bed columns. They reported that the most favourably sorbed metal, Pb^{2+} , occupied most of the adsorption sites on the sawdust in the column. Therefore, less favourably sorbed metals such as Cu^{2+} and Zn^{2+} showed more rapid breakthrough. Also, the slopes of the Zn^{2+} and Cu^{2+} curves were steeper than that of the Pb^{2+} curve.

The removal of chromium from wastewater using Mango seed shell activated carbon (MSSAC) in adsorption columns has not been reported. Chromium exists usually in both trivalent and hexavalent forms in aqueous systems. The two oxidation states of chromium have different chemical, biological and environmental characteristics (WHO, 1988). Chromium (III) is relatively insoluble and is required by microorganisms in small quantities as an essentially trace metal nutrient (Saner, 1980). While Chromium (VI) is of great public health concern because of its toxic effects such as dermatitis, ulceration of the skin, liver and kidney damage in animals and humans. The effective removal of chromium from industrial wastewaters is therefore of great importance. The main sources of chromium wastewater are the industries such as paints and pigments, leather tanning, textile dyeing and electroplating operations.

Benue state is the major producer of mangos in Nigeria and during the harvesting period, which is usually between March and July (depending on the variety), the wastes residue (seeds) which litter the surroundings is of

concern to environmental pollution, especially as it relates to solid waste management. When the two mango processing industries under construction become operational, the rate of generation of this waste will increase, likewise the cost of management, particularly at the company sites. Considering the large quantity of mango seeds generated in this area and their low cost, it is reasonable to study this residue as a possible precursor for the production of activated carbon for wastewater treatment. Elemental analysis of Mango Seed Shells revealed high carbon content (Table 1) implying that it is a good material for conversion to activated carbon. The objective of this paper therefore is to investigate the performance of mango seed shell activated carbons on the removal of chromium in fixed bed down flow columns for the purpose of converting the waste to wealth.

MATERIALS AND METHODS

Chur-kpeve, a local variety of mango (*Magnifera indica*) was used for the production of activated carbons. The mango seeds were sourced from Gboko, Ushongu and Makurdi Local Government Areas of Benue State, air dried and broken to obtain the shells. The shells were thoroughly washed with water to remove any dirt, air dried and cut into sizes of 2 - 4 cm for chemical activation by impregnating with $ZnCl_2$ solution. The $ZnCl_2$ solution was prepared by dissolving 50 g anhydrous zinc chloride in 200 mL of distilled water. 100 g of the shells were mixed with the solution, boiled at 100°C for 30 minutes, and cooled in air to soak for 24 hours. The activated shells were then air-dried. Mixing 50 g of $ZnCl_2$ with 100 g of mango seed shells represented impregnation ratio of 1:2 by mass (i.e. one part of anhydrous $ZnCl_2$ to 2 parts of mango seed shells). Impregnation ratio of 1:3 was also investigated. The $ZnCl_2$ activated shells were carbonized in a muffle furnace at 500°C for 60 minutes to produce *Local* 1:2 and *Local* 1:3 activated carbons. Details of the preparation and characterisation of the Mango seed shell activated carbons have been reported elsewhere (Akpen, 2011; Akpen *et al.*, 2011, Akpen *et al.*, 2014).

Table 1 : Elemental composition of mango seed shells

S/N	Mango Variety	Sulphur (%)	Carbon (%)	Nitrogen (%)
1	Local (Chur-kpev)	1.4 x 10 ⁻⁶	43.05	0.67
2	Dausha	1.6 x 10 ⁻⁶	43.63	0.89

Source: [Akpen, 2012]

Adsorption columns were constructed with Local 1:2 variation of the mango seed shell activated carbon (of particle size 150-300 µm) as adsorbent to exploit their suitability for industrial application in adsorption columns. The contact time and column height were used as variables. Column experiments were performed in a plastic column of 40 mm diameter. A known amount (in grammes) of carbon under study was carefully transferred into the plastic column packed with cotton wool at the bottom to prevent the escape of the carbon particles with the effluent. The water sample to be treated was poured in a 4 litre capacity plastic container fitted with a tap at the bottom to which a 16 mm plastic pipe was connected and was led to the top of the column through an elbow. Effluents through the plastic column were collected in a 1000 ml beaker through another 16 mm plastic pipe with a connected tap to regulate the flow. In order to yield different bed heights, 15 g, 18 g and 21 g of MSSAC were added to produce 3 cm, 3.5 cm and 4 cm, respectively. Before application of the water to be treated to the columns, distilled water was passed through the columns to ensure close packing of the carbon particles. The flow through the columns was by gravity. Efforts were made to keep the pressure head at the desired level above the carbon in the column for all the experimental runs. Lot volumes were collected separately at predetermined time intervals and analysed for residual chromium concentration using Inductively Coupled Plasma (ICP).

Data Analysis

The result obtained from the column studies was used to plot breakthrough curves. BDST

models were also developed at breakthrough concentrations of 10 %, 40 %, 50 %, 60 %, 80 % and 90 % respectively. The validity of the models was verified using Chi-square (χ^2) test according to equation 5:

$$\chi^2 = \frac{(O - E)^2}{E} \quad (5)$$

Where, O is the observed service time of the column and E is the predicted service time of the column using the developed BDST models.

RESULTS AND DISCUSSION

Effect of time and bed height on the removal of chromium from wastewater

The effect of time on the effluent chromium concentration and percentage chromium adsorbed at the bed height of 3cm is shown in Fig.1. The fig. indicates a gradual decrease in the percentage adsorbed with time until the percentage adsorbed became zero at about 72.33 hours. A similar trend was observed using bed heights of 3.5 cm and 4 cm respectively (see Figs. 2 and 3). The observed trend shows that as the adsorption process continued in the column, the active sites of the adsorbent were gradually saturated leading to lesser amount of chromium adsorbed with time.

With respect to bed height, it was observed that percentage chromium removal was significantly affected by bed height (see Figs 1-3). At operation time of five (5) hours for instance, the percentage adsorbed increased from 52.3 % to 69.1 %, when the bed height increased from 3 cm to 4 cm. This increase in percentage adsorbed is expected because the surface area available for adsorption increased with increase in the amount of the adsorbent because more binding sites were available for sorption with increase in bed height.

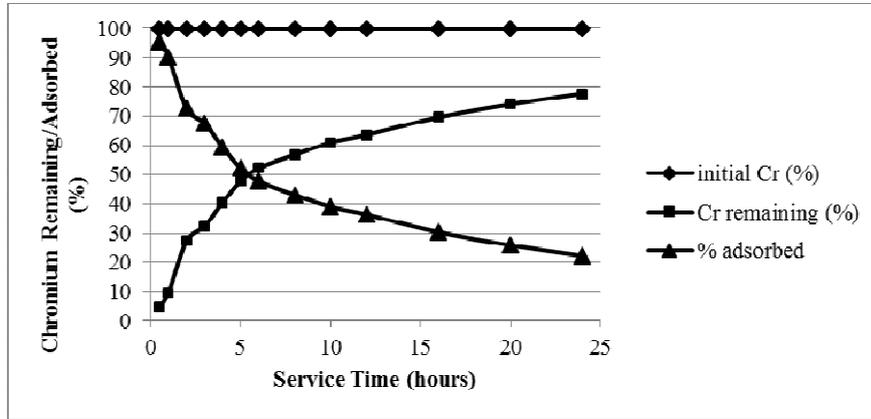


Fig. 1: Effect of time on chromium adsorption at bed height of 3 cm
 (Carbon =Local 1:2, initial chromium=3.151 mg/L)

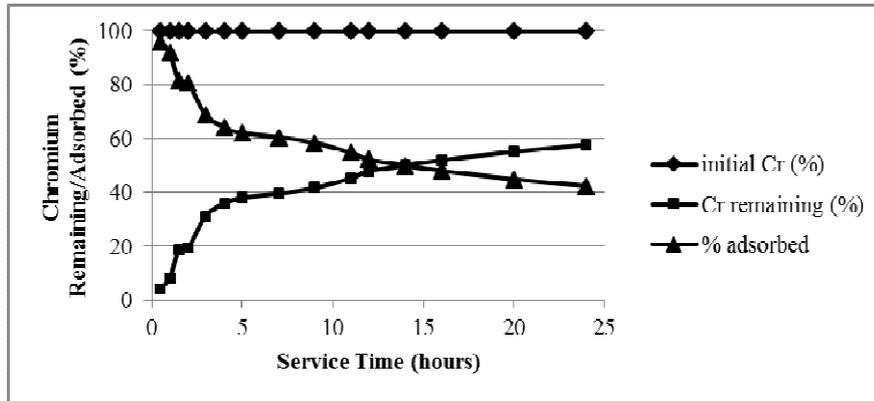


Fig. 2: Effect of time on chromium adsorption at bed height of 3.5cm
 (carbon =Local 1:2, initial chromium=3.151 mg/L)

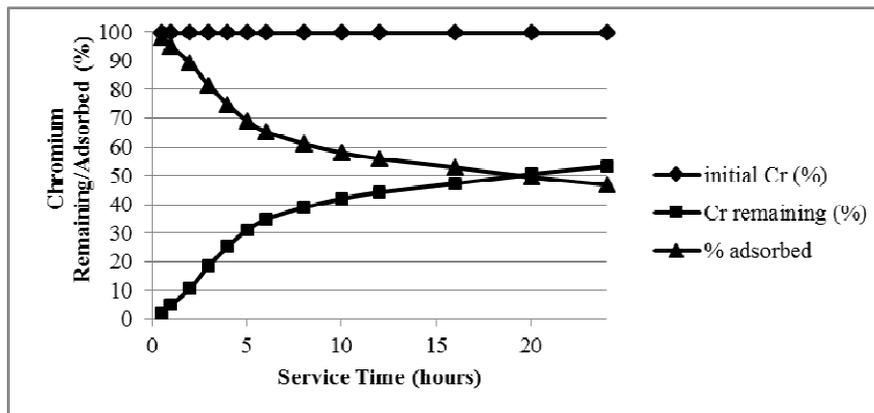


Fig. 3 :Effect of time on chromium adsorption at bed height of 4 cm
 (Carbon = Local 1:2, initial chromium = 3.151 mg/L)

Column breakthrough curves

The most important criterion in the design of fixed bed adsorption systems is the prediction of column breakthrough, which determines the operating life-span of the bed and regeneration times. The adsorption breakthrough curves obtained by varying the bed heights from 3 to 4 cm at constant pressure head of 28 cm and initial chromium concentration of 3.151 mg/L are given in Figs.4-6. The curves were obtained by plotting the dimensionless concentrations, C_t/C_o against the service times. These breakthrough curves deviate from the characteristic S-shape for most dynamic adsorption studies in water and waste water treatment. This is probably because 100% removal of chromium was not attained even at the commencement of the column operation which may be due to the high initial concentration of chromium adopted in this study. Mazumder *et al* (2011) reported a similar deviation on the adsorption of chromium by activated alumina.

The breakthrough and exhaustion points were assumed as 10 % (0.3151 mg/L) and 90 % (2.8359 mg/L) of the initial influent concentration respectively. As shown in the figures, the exhaustion time increased with increasing bed height, from 44.17-290.45 hours as more adsorption sites available for sorption, also resulted in a broadened mass transfer zone. As the feed solution continued to pass through the adsorbent bed, the carbon adsorption sites continued to reduce, and this lead to increase in the concentration of chromium in the effluent with time. At the bed depth 3 cm for instance, breakthrough and exhaustion points were reached after 1.022 hours and 44.17 hours respectively.

BDST plots

The breakthrough data obtained from the column studies was examined using the BDST model (Equation 1). The plots of service time (t) against bed depth (H) at 10 % and 90 % breakthroughs are shown in Figs.7 and 8 respectively. The BDST models at various breakthrough concentrations and their corresponding constants are listed in Table 2. The service time

percentage increased, the BDST rate constant, K_a decreased while the adsorption capacity of the bed per unit bed volume, N_o , and the critical bed depth, H_o , increased. The high values of the coefficient of determination, R^2 (0.804 and 0.916 respectively at 10 % and 90 % breakthrough for instance), indicated the validity of BDST model for the present system.

To confirm the validity of the developed BDST models, a chi-square test was performed (see Table 4). It is obvious from Table 4 that all the models (except 80 % and 90% breakthrough models) were reliable for the prediction of the columns performance given that, the calculated χ^2 values were less than the tabulated values both at 5 % and 1 % levels of significance.

As mentioned earlier, the BDST model constants can be helpful to scale up the process for other flow rates and concentrations without further experimental run (Tables 5 and 6). The use of BDST model in this way provides a realistic description of the adsorption of chromium by MSSAC and the empirical data can be adopted in predicting the bed depth or service time for a specified set of influent characteristics for the scale-up of the adsorption column thus saving time, money and energy that would have been expended in carrying out experiments. The use of BDST plots for column design is illustrated in the next section.

Design of prototype adsorption column

As an illustration, we adopt the 10 % breakthrough concentration BDST model for the design. Its important to note that the breakthrough concentration depends on the desired effluent quality and may be fixed by a pollution control regulation authority such as National Environmental Standards and Regulatory Agency (NESRA) in Nigeria. Suppose a hypothetical industry discharges 1000 m³/8-hour day of wastewater containing 3.151 mg/L of chromium which has to be treated before discharge into a receiving stream. The design is accomplished as follows:

The established BDST model for 10% break

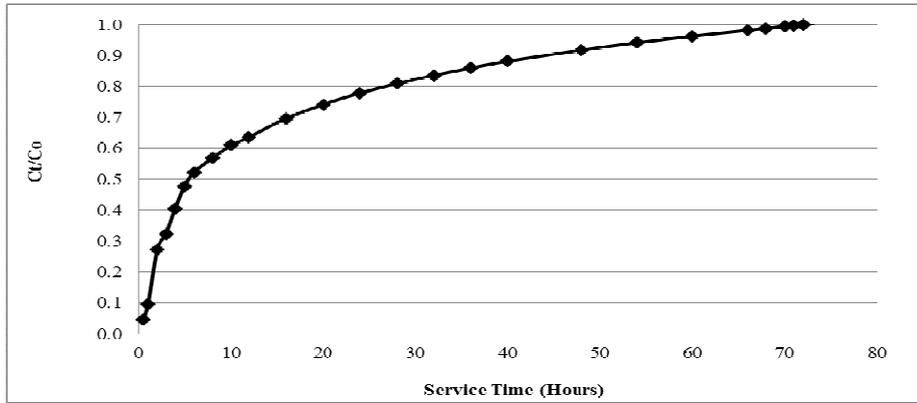


Fig. 4 : Breakthrough curve for chromium adsorption on MSSAC columns with composite wastewater at 3 cm bed height (initial chromium concentration= 3.151 mg/L, pH=6.5, pressure head=28 cm)

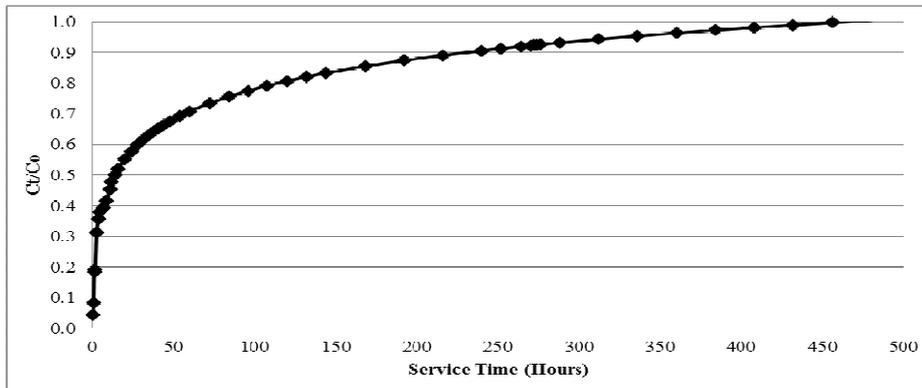


Fig. 5: Breakthrough curve for chromium adsorption on MSSAC columns with composite wastewater at 3.5 cm bed height (initial chromium concentration= 3.151 mg/L, pH=6.5, pressure head=28 cm)

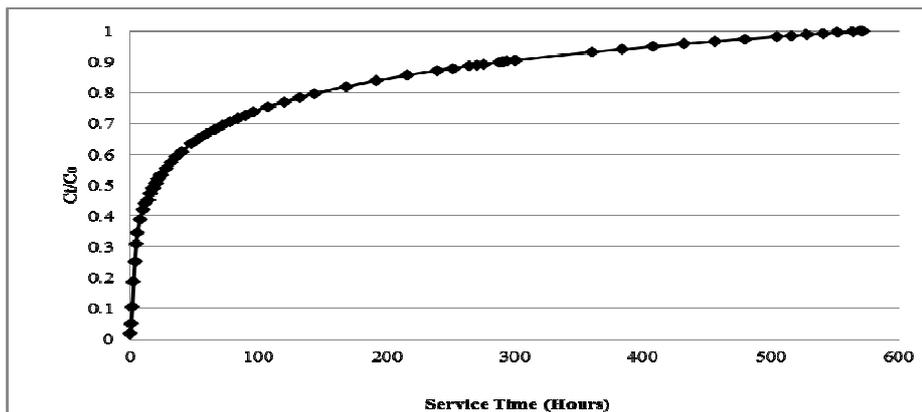


Fig. 6 : Breakthrough curve for chromium adsorption on MSSAC columns with composite wastewater at 4cm bed height (initial chromium concentration= 3.151 mg/L, pH=6.5, pressure head=28 cm)

Table 2 : BDST model parameters

%	BDST model	N_0 (mg/L)	k_a (L/mg.h)	Flow rate (L/h)	H_0 (m)	R^2
10	$t=0.883H - 1.752$	3.564	0.398	1.61	1.984	0.804
40	$t=4.9H - 10.38$	19.78	0.0124	1.61	2.118	0.938
50	$t=13.81H - 35.41$	55.743	0	1.61	0	0.983
60	$t= 28.57H - 74.71$	115.321	-0.0017	1.61	2.649	0.968
80	$t=121.5H - 328.6$	490.426	-0.0013	1.61	2.785	0.936
90	$t= 246.2H - 673.1$	993.77	-0.0010	1.61	2.832	0.916

Table 3: Observed and predicted service times at various breakthrough concentrations

H (m)	10 %		40 %		50 %		60 %		80%		90 %	
	Obs.	Pred.	Obs.	Pred.	Obs.	Pred.	Obs.	Pred.	Obs.	Pred.	Obs.	Pred.
3	1.022	0.017	3.95	4.32	5.51	6.02	9.52	11	26.81	35.9	44.17	65.5
3.5	1.086	0.048	7.49	6.77	13.99	12.93	28.24	25.29	114.95	96.65	231.8	188.6
4	1.905	0.009	8.85	9.22	19.33	19.83	38.09	39.57	148.33	157.4	290.45	311.7

Legend: Obs. = Observed service time

Pred. = Predicted service time (using the developed BDST models)

Table 4: Chi-Square test results on various breakthrough concentrations

H (m)	10 %	40 %	50 %	60 %	80 %	90 %
3	0.017	0.032	0.043	0.199	2.302	6.95
3.5	0.048	0.077	0.087	0.345	3.465	9.9
4	0.009	0.015	0.013	0.055	0.522	1.45
χ^2	0.074	0.124	0.143	0.599	6.289	18.3

NB (i) Degree of freedom = 3-1 = 2

(ii) Chi-square values from Frank and Altheon (1994) are 5.99 and 9.21 at 5% and 1% significant levels respectively.

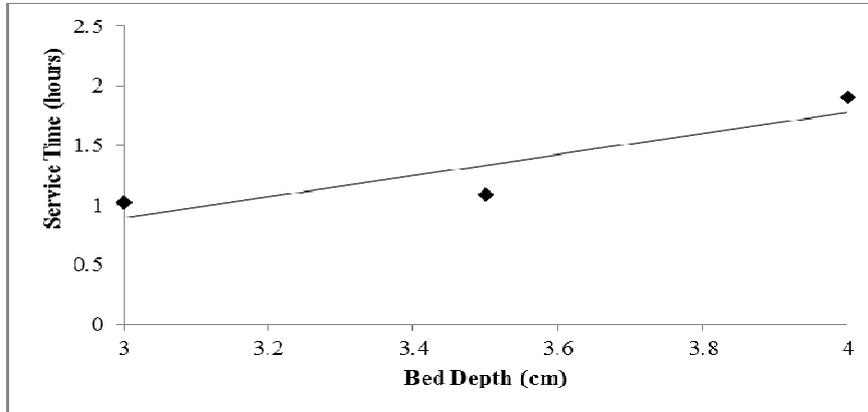


Fig. 7 : BDST plots at 10% breakthrough

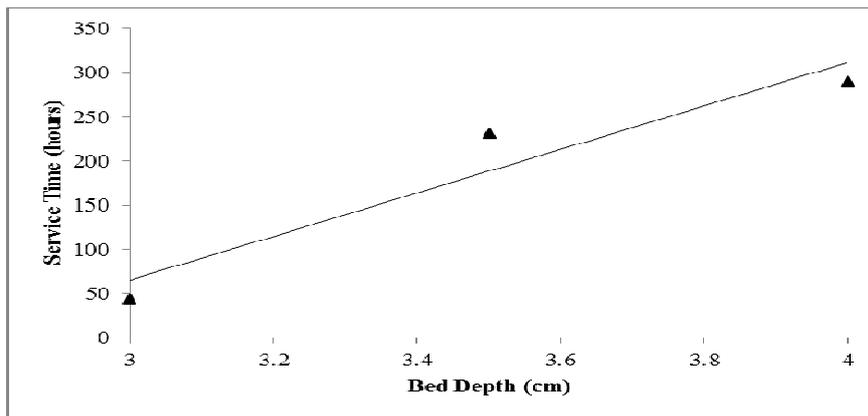


Fig. 8 : BDST plots at 90 % breakthrough

through from test column is (Table 2):
 $t=0.883H-1.752$

The diameter of the test column is 4cm, therefore for a flow rate of 1.61 L/h, the hydraulic loading rate =

$$\frac{Q}{A} = \frac{1.61 \times 10^{-3} \times 4}{3.142 \times 0.04^2} \text{ (m}^3\text{/m}^2\text{/h)}$$

$$= 1.281 \text{ (m}^3\text{/m}^2\text{/h)}$$

Assume design column diameter of 1000mm, then hydraulic loading rate for the flow rate of

$$125\text{m}^3\text{/h (=1000m}^3\text{/8-hour day)} = 159.134\text{(m}^3\text{/m}^2\text{/h)}$$

The change in flow rate will cause a change in slope.

$$\text{The new slope, } a^1 = \frac{1.281}{159.134} \times 0.883 = 0.007108$$

Thus, the BDST model for the design column is:

Table 5 : Predicted BDST constants for other flow rates

Breakthrough (%)	Q ¹ (L/h)	a	a ¹	b=b ¹	t _c (h)
10	0.5	0.833	2.844	-1.752	9.624
	1	0.833	1.422	-1.752	3.936
	2	0.833	0.711	-1.752	1.092
	3	0.833	0.474	-1.752	0.144
	5	0.833	0.284	-1.752	-0.616
90	0.5	246.2	792.764	-673.1	2497.956
	1	246.2	396.382	-673.1	912.428
	2	246.2	198.191	-673.1	119.664
	3	246.2	132.127	-673.1	-144.592
	5	246.2	79.276	-673.1	-355.996

Table 6: Predicted BDST constants for other initial chromium concentrations

Breakthrough (%)	C ¹ (mg/L)	a	a ¹	b	b ¹	t _c (h)
10	1.2	0.833	2.319	-1.752	-4.60	4.676
	1.5	0.833	1.855	-1.752	-3.68	3.74
	2	0.833	1.391	-1.752	-2.760	2.804
	3	0.833	0.927	-1.752	-1.840	1.868
	5	0.833	0.556	-1.752	-1.104	1.12
	7	0.833	0.397	-1.752	-0.789	0.799
	90	1.2	246.2	646.48	-673.1	-1767.47
1.5		246.2	517.184	-673.1	-1413.96	654.777
2		246.2	387.888	-673.1	-1060.47	491.083
3		246.2	258.592	-673.1	-706.979	327.389
5		246.2	155.155	-673.1	-424.188	196.432
7		246.2	110.825	-673.1	-302.991	140.309

$$t = 0.007108H - 1.752$$

$$\text{But } t = 8 \text{ h}$$

$$\text{Therefore, } H = \frac{(8 + 1.752)}{0.007108} = 1372 \text{ cm} = 13.72 \text{ m.}$$

Hence the specifications of the design column are :

$$\text{Bed depth} = 14 \text{ m}$$

$$\text{Diameter} = 1000 \text{ mm}$$

$$\text{Hydraulic loading rate} = 159.134 \text{ m}^3/\text{m}^2/\text{h}$$

Service time = 8 hours

It should be borne in mind that the data used in constructing BDST models herein, were obtained using simulated solutions instead of live wastewater. Impurities in industrial effluents could interfere with the adsorption of chromium and consequently result in an inaccurate prediction in treatment of industrial effluents. Hence the need for pilot studies on live wastewater.

CONCLUSION

The effects of time and carbon height on chromium removal from water were investigated in fixed down flow adsorption columns. There was a gradual decrease in the percentage of chromium adsorbed with time. The percentage adsorbed became zero at 72.33, 467.5 and 572 hours respectively for 3cm, 3.5cm and 4cm carbon heights. These results showed that chromium removal was significantly affected by both carbon height and service time. The breakthrough curves for chromium adsorption deviated from the characteristic S-curve for most dynamic studies. This is because 100 % chromium removal was not attained even at the commencement of the experiment when all the adsorption sites were available due to the high initial concentration of the adsorbate adopted in this study. The Hutchins (1973) BDST model was used to study the columns performance at different breakthrough concentrations. The BDST rate constant k_a , decreased, while adsorption capacity of the bed per unit volume, N_0 , and the critical bed depth, H_0 , increased with increase in breakthrough percentage. The BDST equation obtained at flow rate of 1.61 L/hr and influent chromium concentration of 3.151 mg/L was used to predict the adsorbent performance at other flow rates and initial chromium concentrations. It was observed that for a given breakthrough concentration, the service time decreased with increase in both flow rate and initial concentration. Chi-square test performed to verify the validity of the developed BDST models revealed that the models predicted the columns performance better at lower

breakthrough concentrations. It was recommended that MSSAC can be used for the removal of chromium from wastewater in adsorption columns; however there is need for pilot studies on live wastewater to investigate the effect of interference.

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