

ASSESSMENT OF RADIOLOGICAL HAZARD INDICES FROM SURFACE SOIL TO INDIVIDUALS FROM MAJOR MARKETS AT SAGAMU OGUN STATE, NIGERIA

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ABSTRACT

Assessment of radioactivity concentrations of ^{40}K , ^{238}U and ^{232}Th in surface soil samples of major markets at Sagamu, Ogun State in Nigeria had been determined by gamma spectrometry using Sodium Iodide Thallium doped NaI (TI) detector coupled with a pre-amplifier base to a multiple channel analyzer (MCA). 10 samples each were collected from three major markets: Falawo, Awolowo and Sabo. Highest radioactivity concentrations of ^{40}K , and ^{238}U were obtained from Falawo market soil samples with values 1274.26 ± 4.26 Bq/kg and 40.72 ± 3.12 Bq/kg respectively while that of ^{232}Th was obtained from Sabo soil sample with value 115.62 ± 16.39 Bq/kg. The mean external hazard index (H_{ex}) and mean internal hazard index (H_{in}) for all the soil samples from Falawo market were calculated to be 0.616 Bq/kg and 0.691 Bq/kg respectively, and that of Awolowo market were 0.566 Bq/kg and 0.634 Bq/kg respectively. Also for Sabo market the mean values were calculated to be 0.594 Bq/kg and 0.658 Bq/kg respectively. All the values obtained were less than 1.0 Bq/kg as recommended for safety and therefore have no negative radiological health implication to the people within the markets and their environs.

Keywords: Radioactivity concentration, radionuclide, hazard index, surface soil.

INTRODUCTION

All living things are continuously exposed to ionizing radiation from Naturally Occurring Radioactive Materials (NORM). Naturally occurring radioactive materials (NORM) existing in soil could pose potential health physics risk (Wilson, 1993), especially if assisted by natural processes such as weathering deposition and wind erosion (Elles *et al.*, 1997). The artificial sources are largely due to medical and industrial activities. Studies on radiation levels and radionuclide distribution in the environment provide vital radiological baseline information. Such information is essential in understanding human exposure from natural and man-made sources of radiation and necessary in establishing rules and regulations relating to radiation protection (Quindos *et al.*, 1994).

The Earth is naturally radioactive, and about 90% of human radiation exposure arises from natural sources such as cosmic radiation, exposure to radon gas and terrestrial radionuclides (Lee *et al.*, 2004). There are many sources of radiation and radioactivity in the environment. The earth and atmosphere contain varied levels of naturally sourced radionuclides such as ^{238}U and ^{232}Th decay chains as well as singly occurring types such as ^{40}K . Soil features, geological formations, and human activities related to radiation and radioactivity are important factors enhancing the background levels of natural radiation (Colmerero Sujo *et al.*, 2004). The continuity in increasing these radionuclides in the environment may be attributed to several factors such as the successive utilization of phosphate fertilizer, burning of fossil fuels (crude oil and coal), mining and milling operations,

and building materials. Ingesting and inhaling such levels of radionuclides contribute significantly to the radiation dose that people receive (Martínez, 1989). In addition, Mining and milling of both nuclear and non-nuclear materials may cause significant environmental and occupational radiological impacts, and, typically, NORM in commercial and industrial products has the potential to expose workers and members of the public to some fraction of the recommended annual radiation exposure limit. The main external source of irradiation to the human body is represented by the gamma radiation emitted by naturally occurring radioisotopes, also called terrestrial environmental radiation. These radioisotopes, such as ^{40}K and the radionuclides from the ^{232}Th and ^{238}U series and their decay products, exist at trace levels in all ground formations. Therefore, natural environmental radioactivity and the associated external exposure due to gamma radiation depend primarily on the geological and geographical conditions, and appear at different levels in the soils of each different geological region (UNSCEAR 1993,2000 Reports). However, it has been observed that the type and concentration vary considerably depending on the soil type. The effects of the radiation emitted by different radionuclides depend on the overlying soil material (thickness and type), its chelating agents and physio-chemical properties (Believermis *et al.*, 2009). Investigation has shown that natural radioactivity and the associated exposure due to gamma radiation (from radionuclides) depend primarily on geology (soil type).

The interaction of ionizing radiation with human body leads to various biological effects which may later show up as clinical symptoms (ICRP, 1990).

The study of the radionuclide concentration levels in soil and water samples in Eagle, Atlas and rock cement companies in Port Harcourt was carried out by Awwiri (2005) soil and water samples collected from the respective premises were analyzed using the gamma -ray spectrometry. The average absorbed dose rates of the soil samples were 49.27nGy/h, 45.21nGy/h and 42.33nGy/h for Eagle, Atlas and Rock cements respectively while the water values were 22.16nGy/h, 20.75nGy/h and 19.37nGy/h for the respective companies. Mean dose rate equivalents of 0.18mSv/y and 0.39mSv/y were obtained for the water and soil samples. These results are lower than the International Commission on Radiological Protection (ICRP, 1992) maximum permitted limit and therefore, have no significant radiological health burden on the environment and the populace.

The objective of this work however is to determine the radioactivity level of naturally occurring ^{40}K , ^{238}U and ^{232}Th in surface soil samples from major markets at Sagamu, in Ogun State of Nigeria. The measured radioactivity concentrations would thereafter translate to the calculation of the internal and external hazard indices in order to assess the radiological implication to the people within the markets in selling and buying processes, and those involved in other activities.

MATERIALS AND METHODS

Surface soil samples were collected randomly at different points within three major markets: Falawo, Awolowo and Sabo at Sagamu in Ogun State, southwest of Nigeria. Ten (10) samples each were collected from the markets at the same depth from 0.0cm to 20.0cm. A total of thirty soil samples were collected. They were put in different containers and taken to the laboratory to dry for about seventy two hours (72.0hrs) under laboratory temperature of about 27°C and relative humidity of about 70% (IAEA, 1989). Each dried soil sample was crushed and sieved using a 2 mm mesh screen. The dried samples were then packed 150.0 g by mass in labeled cylindrical plastic containers of uniform base diameter of 5.0 cm which could sit on the 7.6 cm by 7.6 cm NaI (TI) detector. The plastic containers were tightly covered, sealed and left for 28 days prior to counting, for attainment of secular equilibrium between ²³⁸U and ²³²Th and their respective progenies (Papaefthymiou, 2007).

The method of gamma spectrometry was adopted for the analysis of the samples collected in order to obtain data on ⁴⁰K, ²³⁸U and ²³²Th. The spectrometer used was a Canberra lead shielded 7.6cm x 7.6cm NaI (TI) detector coupled to a multichannel analyzer (MCA) through a preamplifier base. The system was calibrated using standard point sources of gamma emitting isotopes. The resolution of the detector is about 10% at 0.662MeV of ¹³⁷Cs. According to Jibiri and Farai (1998) the value is good enough for NaI detector to distinguish the gamma ray energies of most radionuclides in the samples. For the analysis of ⁴⁰K, ²³⁸U and ²³²Th, the photo peak regions of ⁴⁰K (1.46 MeV), ²¹⁴Pb (1.76 MeV) and ²⁰⁸Tl (2.615 MeV) were respectively used.

The cylindrical plastic containers holding the samples were put to sit on the high geometry 7.6cm x 7.6cm NaI (TI) detector. High level shielding against the environmental background radiation was achieved by counting in a Canberra 10cm thick lead castle. The counting of each sample was done for 10.0 hours because of suspected low activities of the radionuclides in the samples. The areas under the photo-peaks of ⁴⁰K, ²³⁸U and ²³²Th were computed using the Multichannel Analyzer system.

The concentrations of the radionuclides were calculated based on the measured efficiency of the detector and the net count rate under each photopeak over a period of 10.0 hours using equation 1.0 (IAEA, 1989)

$$A = \frac{N(E_{\gamma})}{\epsilon(E_{\gamma})IMt_c} \tag{1.0}$$

Where:

N(E_γ) = Net peak area of the radionuclide of interest

ε(E_γ) = Efficiency of the detector for the γ- energy of interest

I_γ = Intensity per decay for the γ- energy of interest

M = Mass of the sample

t_c = Total counting time in seconds (36000s)

External hazard index (H_{ex}) as expressed in equation 2.0 (Hamzah, 2008) and internal hazard

$$H_{ex} = \frac{C_k}{4810} + \frac{C_{Th}}{259} + \frac{C_{Ra}}{370} \leq 1 \tag{2.0}$$

index (H_{in}) as expressed in equation 3.0 were used as radiological indicators to estimate the radiological implications of the natural radionuclides in the surface soil on the people in the market.

$$H_{in} = \frac{C_k}{4810} + \frac{C_{Th}}{259} + \frac{C_{Ra}}{185} \leq 1 \tag{3.0}$$

C_k, C_{Th} and C_{Ra} are the radioactivity concentrations of ⁴⁰K, ²³²Th and ²²⁶Ra respectively.

RESULTS AND DISCUSSION

The highest radioactivity concentration of ⁴⁰K from Falawo market soil sample was obtained to be 1274.26 ± 4.26 Bq/kg and the lowest was 785.49 ± 2.23 Bq/kg. For ²³⁸U, the highest radioactivity concentration value was 40.72 ± 3.12 Bq/kg and the lowest value was obtained to be 15.48 ± 1.76 Bq/kg. Moreso, ²³²Th had the highest radioactivity concentration value of 112.83 ± 15.35 Bq/kg and lowest value of 65.31 ± 7.24 Bq/kg as shown in table 1.0. Concerning all the radionuclides, ⁴⁰K had the highest radioactivity concentration value while the lowest was from ²³⁸U. No artificial radionuclide was detected in all the samples. The highest external hazard index (H_{ex}) was obtained to be 0.746 Bq/kg while the lowest was 0.500 Bq/kg. The mean value was calculated to be 0.616 Bq/kg. Furthermore, the highest internal hazard index (H_{in}) was obtained to be 0.823 Bq/kg while the lowest was 0.571 Bq/kg. The mean value was calculated to be 0.691 Bq/kg.

For Awolowo market soil samples, the highest radioactivity concentration of ⁴⁰K was obtained to be 1064.84 ± 5.36 Bq/kg and the lowest was 568.73 ± 2.86 Bq/kg. For ²³⁸U, the highest radioactivity concentration value was 40.26 ± 2.84 Bq/kg and the lowest value was obtained to be 14.08 ± 1.86 Bq/kg. Also, ²³²Th had the highest radioactivity concentration value of 110.02 ± 14.85 Bq/kg and lowest value of 48.35 ± 6.37 Bq/kg as shown in table 2.0. Concerning all the radionuclides, ⁴⁰K has the highest radioactivity concentration value while the lowest was from ²³⁸U as shown in table 2.0. No artificial radionuclide was detected in all the samples. The highest external hazard index (H_{ex}) was obtained to be 0.688 Bq/kg while the lowest was 0.438 Bq/kg. The mean value was calculated to be 0.566 Bq/kg. Furthermore, the highest internal hazard index (H_{in}) was obtained to be 0.775 Bq/kg while the lowest was 0.485 Bq/kg. The mean value was calculated to be 0.634 Bq/kg.

Also, for Sabo market soil samples, the highest radioactivity concentration of ⁴⁰K was obtained to be 1124.85 ± 6.58 Bq/kg and the lowest was 748.68 ± 3.01 Bq/kg. For ²³⁸U, the highest radioactivity concentration value was 32.34 ± 3.28 Bq/kg and the lowest value was obtained to be 16.64 ± 2.82 Bq/kg. Also, ²³²Th had the highest radioactivity concentration value of 115.62 ± 16.39 Bq/kg and lowest value of 56.24 ± 8.02 Bq/kg as shown in table 3.0. Concerning all the radionuclides, ⁴⁰K had the highest radioactivity concentration value while the lowest was from ²³⁸U as shown in table 3.0. No artificial radionuclide was detected in all the samples. The highest external hazard index (H_{ex}) was obtained to be 0.735 Bq/kg while the lowest was 0.470 Bq/kg. The mean value was calculated to be 0.594 Bq/kg. Furthermore, the highest internal hazard index (H_{in}) was obtained to be 0.819 Bq/kg while the lowest was 0.521 Bq/kg. The mean value was calculated to be 0.658 Bq/kg. The mean values are less than 1.0 Bq/kg as recommended for safety globally (EC, 1999).

Table 1.0: Radioactivity concentrations, external and internal indices of Falawo soil samples

MARKET	SAMPLE	RADIOACTIVITY CONCENTRATIONS OF SOIL SAMPLES (Bq/kg)			H_{ex} (Bq/kg)	H_{in} (Bq/kg)
		^{40}K	^{238}U	^{232}Th		
FALAWO	C ₁	1274.26 ± 4.26	36.58 ± 2.78	65.31 ± 7.24	0.616	0.715
	C ₂	1025.38 ± 3.96	20.17 ± 2.26	95.24 ± 10.46	0.635	0.690
	C ₃	865.94 ± 2.54	21.35 ± 2.32	85.10 ± 8.37	0.566	0.624
	C ₄	798.16 ± 2.25	15.48 ± 1.76	101.27 ± 11.08	0.599	0.641
	C ₅	984.36 ± 3.26	32.74 ± 2.61	104.18 ± 12.14	0.695	0.784
	C ₆	1092.47 ± 4.05	40.72 ± 3.12	72.95 ± 8.02	0.619	0.729
	C ₇	1126.28 ± 4.15	28.25 ± 2.45	112.83 ± 15.35	0.746	0.823
	C ₈	847.32 ± 2.31	18.59 ± 1.84	92.37 ± 10.22	0.583	0.633
	C ₉	785.49 ± 2.23	26.24 ± 2.40	68.74 ± 7.63	0.500	0.571
	C ₁₀	925.73 ± 3.06	34.62 ± 2.55	82.65 ± 7.98	0.605	0.699

Table 2.0: Radioactivity concentrations, external and internal indices of Awolowo soil samples

MARKET	SAMPLE	RADIOACTIVITY CONCENTRATIONS OF SOIL SAMPLES (Bq/kg)			H_{ex} (Bq/kg)	H_{in} (Bq/kg)
		^{40}K	^{238}U	^{232}Th		
AWOLOWO	D ₁	982.29 ± 3.18	17.35 ± 2.36	48.35 ± 6.37	0.438	0.485
	D ₂	1046.58 ± 4.12	25.27 ± 2.48	85.21 ± 8.57	0.615	0.683
	D ₃	568.73 ± 2.86	15.88 ± 2.02	72.94 ± 9.36	0.443	0.486
	D ₄	845.26 ± 3.05	32.26 ± 2.53	110.02 ± 14.85	0.688	0.775
	D ₅	785.32 ± 2.37	24.42 ± 2.41	98.68 ± 12.07	0.610	0.676
	D ₆	694.38 ± 2.94	26.57 ± 2.43	104.12 ± 13.14	0.618	0.690
	D ₇	925.47 ± 3.07	31.04 ± 2.51	52.73 ± 7.76	0.480	0.564
	D ₈	1064.84 ± 5.36	40.26 ± 2.84	76.16 ± 10.24	0.624	0.733
	D ₉	994.36 ± 3.46	14.08 ± 1.86	68.59 ± 7.92	0.510	0.548
	D ₁₀	1004.18 ± 3.71	22.24 ± 2.08	95.34 ± 11.86	0.637	0.697

Table 3.0: Radioactivity concentrations, external and internal indices of Sabo soil samples

MARKET	SAMPLE	RADIOACTIVITY CONCENTRATIONS OF SOIL SAMPLES (Bq/kg)			H _{ex} (Bq/kg)	H _{in} (Bq/kg)
		⁴⁰ K	²³⁸ U	²³² Th		
SABO	E ₁	898.25 ± 3.42	16.64 ± 2.82	74.37 ± 9.23	0.519	0.564
	E ₂	984.37 ± 3.64	19.78 ± 2.96	89.24 ± 10.31	0.603	0.656
	E ₃	1082.28 ± 6.01	31.15 ± 3.24	110.19 ± 13.82	0.735	0.819
	E ₄	796.43 ± 3.24	24.23 ± 2.99	96.76 ± 11.45	0.605	0.670
	E ₅	824.36 ± 3.26	20.76 ± 2.87	68.15 ± 8.75	0.491	0.547
	E ₆	1124.85 ± 6.58	27.35 ± 3.14	82.76 ± 9.73	0.627	0.701
	E ₇	969.24 ± 3.37	18.92 ± 2.92	56.24 ± 8.02	0.470	0.521
	E ₈	748.68 ± 3.01	22.17 ± 2.76	97.34 ± 12.04	0.591	0.651
	E ₉	894.46 ± 3.41	32.34 ± 3.28	115.62 ± 16.39	0.720	0.807
	E ₁₀	1042.74 ± 5.48	23.86 ± 2.96	77.73 ± 9.35	0.581	0.646

CONCLUSION

The radiological safety of the people within the markets had been assessed and the results obtained show that the mean external and internal hazard indices values are less than 1.0 Bq/kg, the limit recommended by European Commission on Radiation Protection (EC, 1999). This means that the exposure of the people in the markets to the radiation from the surface soils does not pose any negative radiological effect on them and their environs.

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