

# Vegetation Characteristics and Deforestation at Two Mangrove Ecosystems Subjected to Varying Anthropogenic Influences: Case of Mtoni and Dege, Dar es Salaam, Tanzania

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# ABSTRACT

Mangrove ecosystems are subject to overexploitation, pollution, and conversion to other land uses from anthropogenic pressures. To understand the way different species. mangrove respond to the anthropogenic impacts, Mtoni and Dege mangrove ecosystems, varying with degradation levels were compared on vegetation characteristics, deforestation, and abiotic variables. The study adopted the line transect permanent plots method. In each mangrove sampling plot, vegetation characteristics and selected abiotic variables were assessed. Mangrove vegetation characteristics were using tested an independent t-test and a special t-test. Mangrove species diversity was calculated using Shannon-Wiener Index. Relationships of variables were tested using Spearman's rank correlation coefficients. Findings showed that Mtoni had higher mangrove species diversity, richness, and evenness than Dege. Dege showed significantly higher mangrove density, basal area. and regeneration than Mtoni. Stump density was significantly higher at Mtoni than at Dege. In Mtoni, the basal area was significantly negative correlated with both salinity and organic matter. In Dege, basal area and salinity significantly positive were correlated. There was high mangrove degradation at Mtoni compared to Dege. It is recommended that participatory conservation and management interventions be undertaken. Mere protection from further exploitation is sufficient for Dege, while active restoration is recommended for Mtoni.

**Key words**: Mangrove forests – species diversity – sapling regeneration – mangrove deforestation.

# INTRODUCTION

Mangroves refer to a diverse group of salttolerant plants that inhabit the intertidal margins of low-energy coastlines, mudflats, and river banks in tropical and sub-tropical areas (Friess et al. 2019, UNEP 2014). Mangroves form a complex community below the high tide mark on sheltered tropical shores (Diniz et al. 2022, UNEP 2014). They often form a borderline between the oceans and tropical rain forests; they comprise trees and shrubs belonging to 12 genera in 8 different families worldwide. The dominant mangrove genera are Rhizophora, Avicennia, Sonneratia and Bruguiera (Malik et al. 2015). More than 50% of the world's 100,000 km<sup>2</sup> mangrove forests are found in the eastern hemisphere (UNEP 2014, Linneweber 2013).

As highlighted by Webber et al. (2016), mangrove ecosystems are important for millions of people around the world as they support both subsistence and commercial fisheries, and provide many other ecological and socio-economic benefits. These ecosystems also contain high biodiversity of animals and plants and provide opportunities for ecotourism and education-related activities (Thomas et al. 2018). However, large mangrove areas are subject to increasing pressures from a variety of anthropogenic activities such as agriculture, prawn farming, fishing, salt making, waste disposal, settlements, and cutting of mangroves for fuel, timber, and building poles (Branoff 2017). These pressures threaten mangrove ecosystems, and their affiliated ecosystems of coral reefs and sea grasses, the consequences of which include loss of valuable mangrove resources and a reduction mangrove ecosystem in production, unless effective management initiatives are undertaken (Feka and Morrison 2017).

In Tanzania, mangroves occur along the coast from the border with Kenya in the north to that of Mozambique in the south and around many of the islands off the coast (Basha 2018, Mangora 2016). Recent remote sensing data estimates the total coverage of mangroves in Tanzania at 133,500 ha (Basha 2018, Brown 2016) with major species being Avicennia marina, Bruguiera gymnorrhiza, Heritiera **Ceriops** tagal. littoralis. Rhizophora Lumnitzera racemosa, mucronata, Sonneratia alba, Xylocarpus granatum and Xylocarpus mullucensis (Mangora 2016, Mohamed 2004).

Mangrove ecosystems are threatened by the impacts of anthropogenic pressures (Diniz et al. 2022, Basha 2018, Branoff 2017, Feka and Morrison 2017). However, the response of various mangrove species to these pressures in respect of their resilience is scantly investigated. This study, therefore, investigated the phenological and species diversity-based responses of the two mangrove forests (Mtoni and Dege) varying in their anthropogenic influences along the coast of Dar es Salaam to establish the level of interventions required for protection or restoration. Specifically, the study aimed at assessing mangrove species diversity, density, and basal area, examining the extent of mangrove harvesting, and evaluating the relationship of mangrove basal area with the selected abiotic factors. In lieu of these, the study tested the following hypotheses:

i. Mangrove species diversity, density, and basal area are greater in Dege than in Mtoni

- ii. The extent of mangrove harvesting is greater in Mtoni than in Dege
- iii.Mangrove basal area is positively correlated with sediment organic matter.

# MATERIAL AND METHODS

# The study sites

This study was carried out in the Dege and Mtoni mangrove ecosystems, all located along the coast of Dar es Salaam in Tanzania (Figure 1). These sites were selected because they are large mangrove ecosystems subjected to large human populations that potentially threaten their existence. Dege mangrove ecosystem is found at the mouth of Bandarini River, which is a seasonal stream. It is located at latitude 6° 52' S and longitude 39° 28' E. Its distance from Dar es Salaam city center is about 60 km.

Dege mangrove ecosystem is estimated to cover 245.0 ha, with its main mangrove species being Ceriops tagal, Rhizophora mucronata, Sonneratia alba, Avicennia *Xylocarpus* granatum marina, and gymnorrhiza (Mtanga and Bruguiera Machiwa 2008). The inhabitants of Dege village depend largely upon marine and coastal resources. Their settlement is located just 3.3 km from the mangrove forest, posing great threats to the forest by harvesting mangroves for building poles, firewood, and boat construction (Kristensen et al. 2011).

Mtoni mangrove ecosystem is found along the two creeks, Kizinga and Mzinga, which flow into the Mtoni estuary near Mbagala. It is located between the latitude 6° 45' S, and longitude 39° 41' E, at a distance of about 20 km south of Dar es Salaam city center. This study was conducted along the Mzinga creek, which is approximately 1.5 km across.



Figure 1. The coast of Dar es Salaam showing the location of the study sites. *Source*: University of Dodoma GIS Laboratory, 2022.

Mtoni mangrove forest is estimated to cover 378.4 ha, with dominant mangrove species being *Sonneratia alba, Rhizophora mucronata, Avicennia marina,* and *Ceriops tagal* (Mgaya *et al.* 2004). Mtoni mangrove ecosystem is impacted by anthropogenic activities through cutting down trees for fuel wood and building poles, and domestic and industrial pollution (Mihale *et al.* 2021).

# Characteristics of human populations near the study sites

Dar es Salaam city is situated on Tanzania's east coast at latitudes 6°45' S and 7°25' S and longitudes 39°E and 39°55' E. It shares borders with the Coast Region to the north, west, and south, as well as the Indian Ocean to the east. Dar es Salaam city has a total land area of 1,630.7 km<sup>2</sup> (Moshi *et al.* 2018). The coastal plain and the inland plateau make up the two distinct geographies of Dar es Salaam city. Site elevations range from less than 5 meters above sea level (asl) in the lowlands along the coast to 60-150 meters above sea level (asl). The annual rainfall in the Dar es Salaam city is above 1,000 mm, with two distinct rainy seasons; March to May and October to December (Manara 2020). The city's average monthly relative humidity ranges from 72% in January to 82% in April, with a mean annual temperature range of  $30.8^{\circ}$  C to  $21.3^{\circ}$  C (Ndetto and Matzarakis 2013).

Dar es Salaam city is estimated to have a population of 7 million people, making up 8% of the national population (www.worlpopulationreview.com). Over 70% of the country's gross domestic product (GDP) is produced in Dar es Salaam city, and about 70% of its residents rely on fuelwood like firewood and charcoal for their energy needs (Lyimo 2006). Due to the profitability of the wood industry, illegal logging occurs. The future of Dar es Salaam's natural environment is in danger due to the mangroves being destroyed by deforestation and soil erosion. People living near the Dar es Salaam coast pose socioeconomic threats to mangroves by cutting down mangroves for fuel used in making salt, burning lime, and



smoking fish (Maseta *et al.* 2021). Along the coast of Dar es Salaam city, human populations are also responsible for the removal of mangroves for settlements and poles, as well as by trampling (Mabula *et al.* 2016).

#### Research design and sampling procedures

The study adopted the line transect permanent plots sampling design (English et al. 1997). In each study site, transects were established perpendicular to vegetation zones, covering the whole area from the lower to the upper parts of the creek passing through the mangrove forests. Along each transect, 10m x 10m sampling plots with a distance of 20m between plots, were established. At Mtoni, 6 transects were set, while at Dege 7 transects were set. In each plots transect. 4-8 were established depending upon the length of transect. A total of 79 sampling plots (31 for Mtoni and 48 for Dege), as the sample size (n), was established. The Global Positioning System (GPS) readings were recorded for all plots.

# Quantification of mangrove vegetation characteristics

In each sampling plot, all mangroves were identified to species level and counted according to their maturity categories. The girth at breast height (GBH, standardized as 1.3m above ground) of trees (>8cm in girth) and saplings (<8cm in girth, > 1m in height) were measured using a tape. Seedlings (<1m in height) were also identified to species level and counted. All stumps of cut trees were counted and the girth at the top was measured. If species of stumps could be identified, this was also recorded.

# Measurement of the abiotic factors

Abiotic factors were measured in each study site in the same sampling plots that were selected for the quantification of mangrove species.

*Sediment interstitial salinity:* Interstitial water was drawn from the sediment using a 20-ml syringe. A few drops were then placed

in a refractometer to measure salinity in parts per thousand (ppt).

*Percent saturation capacity:* One random sediment sample was collected in each sampling plot. Approximately 10g from each sample was completely saturated, weighed, placed in a porcelain crucible, and dried in an oven at 105 °C to a constant weight. The percentage loss in weight was then calculated to give percent saturation capacity.

Sediment organic matter content: The ignition method was used to determine the sediment organic matter content of each sample that had been dried for obtaining percent saturation capacity. Sediment samples were burnt at 500°C for 4 hours and then cooled in desiccators. Organic matter content was obtained by calculating the percent loss in weight.

# Data analysis

Data from the two study sites were assumed to be independent, normally distributed, and have a homogeneity of variance. With a guide from the three study hypotheses, these data conditions enabled differences in the mangrove density and the basal area, as well as the level of mangrove harvesting between the two study sites, to be tested using the two-sample/independent t-test with а probability of 0.05, and the degrees of freedom (df) of 77 (i.e.,  $n_1+n_2 - 2$ ). Mangrove species diversity was calculated for each study site using the Shannon-Wiener Index of species diversity, which takes into consideration the number of species as well as the evenness of the abundance of each species (Nolan and Callahan 2006). It was calculated by taking the number of each species, the proportion of each species, and summing up the proportion times the natural logarithms of the proportion for each species as follows:

$$H' = -\sum_{i=1}^{S} p_i \ln p_i$$

Where; H' = species diversity index, s = number of species, and pi = proportion of

individuals of each species belonging to the *i*th species of the total number of individuals.

The difference in diversity indices between the two sites was tested using a special t-test (Zar 1999), with a degrees of freedom (df) of 529 as it considered the number of individual mangrove species in all plots of the two study sites. Correlations between mangrove basal area and mangrove density, and the selected abiotic factors (i.e., sediment organic matter content, percent saturation capacity, and salinity) were tested using Spearman's rank correlation coefficients. This model measured the strength and direction (negative or positive) of the association between the mangrove biotic factors and selected abiotic factors in each study site.

# RESULTS

# Mangrove vegetation characteristics

#### Mangrove species diversity

Based on the Shannon-Weaner Index, mangrove species diversity was significantly greater in Mtoni (0.61) than in Dege (0.50) (special t-test for comparing indices of diversity: t = 2.540, df = 529, 0.01 0.02), with 7 species being observed in Mtoni and 6 in Dege (Table 1). Avicennia marina, Ceriops tagal, Sonneratia alba, Rhizophora mucronata, Bruguiera gymnorrhiza, and Xylocarpus granatum were observed in both forests, while Lumnitzera racemosa was observed in Mtoni only.

Table 1. Species richness, evenness, and<br/>species diversity (Shannon-Weaver<br/>index) of mangroves for Mtoni and Dege<br/>forests along the coast of Dar es Salaam

| Study<br>site | No. of species | Evenness | Species<br>diversity |
|---------------|----------------|----------|----------------------|
| Mtoni         | 7              | 0.72     | 0.61                 |
| Dege          | 6              | 0.64     | 0.50                 |

#### Mangrove basal area

Mangrove basal area was significantly higher in Dege ( $1006cm^2/100-m^2$  plot) than Mtoni ( $555cm^2/100-m^2$  plot) (two-sample ttest: t = 4.476, df = 77, p = 2.60 x  $10^{-5}$ ; Figure 2) at probability of 0.05. In Dege, *Ceriops tagal* had the highest basal area, followed by *Rhizophora mucronata*; while in Mtoni, *Avicennia marina* had the highest basal area followed by *Ceriops tagal* (Figure 3).







Figure 3. Basal area (mean + standard error) of various mangrove species in Mtoni and Dege forests along the coast of Dar es Salaam.

Mangrove stand density (taken as the total number of trees and saplings per unit area) was higher in Dege (251.6 trees/100-m<sup>2</sup> plot) than Mtoni (102.8 trees/100-m<sup>2</sup> plot) (two-sample t-test: t = 2.821, df = 77, p = 0.00061; Figure 4) at probability of 0.05. There was a higher density of all maturity categories at Dege than at Mtoni (Figure 5). In all study sites, *Ceriops tagal* was dominant in both tree and sapling categories, while *Avicennia marina* and *Ceriops tagal* seedlings were dominant in Mtoni (Figure 6) and Dege (Figure 7) respectively.

#### Mangrove regeneration

There was significantly greater regeneration (as indicated by the density of seedlings) at Dege (141.4 seedlings/100-m<sup>2</sup> plot) than Mtoni (36.1 seedlings/100-m<sup>2</sup> plot) (two-sample t-test: t = 3.668, df = 77, p = 0.00045; Figure 5) at probability of 0.05. At Mtoni, *Avicennia marina* had the most numerous seedlings (Figure 6), while at Dege; *Ceriops tagal* had by far the highest density of seedlings (Figure 7).







Figure 5: Total number of trees, saplings, and seedlings (all species combined; mean + standard error) at Mtoni and Dege along the coast of Dar es Salaam



Figure 6. The density of mangrove species (mean + standard error) of various maturity categories at Mtoni forest along the coast of Dar es Salaam



Figure 7. The density of mangrove species (mean + standard error) of various maturity categories at Dege forest along the coast of Dar es Salaam

#### Human pressures on mangroves

Mtoni had significantly higher mangrove stump density than Dege, with 38.7 stumps/100-m<sup>2</sup> and 25.9 stumps/100-m<sup>2</sup> plot respectively (two-sample t-test: t = 2.182, df = 77, p = 0.0322; Figure 8) at probability of indicates This 0.05. that mangrove degradation was high in Dege compared to Mtoni. In both sites, Ceriops tagal had the highest stump density, followed by *Rhizophora* mucronata Dege and at

Avicennia marina at Mtoni (Figure 9). It implies that the preferences for these species (*Ceriops tagal* and *Rhizophora mucronata*) are high compared to other species. However, *Ceriops tagal* had higher stump density in Dege than Mtoni.

#### Selected abiotic factors

The findings for abiotic factors (organic matter, saturation capacity, and interstitial salinity) are shown in Table 2.



Figure 8. Stand stump density (mean + standard error) in Mtoni and Dege forests along the coast of Dar es Salaam



Figure 9. Stump density (mean + standard error) of various mangrove species at Mtoni and Dege forests along the coast of Dar es Salaam

| Table 2. Abiotic | factors as measured | at Mtoni and | Dege forests a | long the coast | of Dar es |
|------------------|---------------------|--------------|----------------|----------------|-----------|
| Salaam           |                     |              |                |                |           |

| A biotic footowa          |      | Mtoni |     |      | Dege |     |
|---------------------------|------|-------|-----|------|------|-----|
| Ablotic factors           | Mean | n     | SE  | Mean | n    | SE  |
| Organic matter (%)        | 6.5  | 31    | 1.6 | 5.1  | 48   | 0.4 |
| Saturation capacity (%)   | 47.6 | 31    | 3.2 | 31.1 | 48   | 1.5 |
| Interstitial salinity (%) | 34.2 | 31    | 0.2 | 32.4 | 48   | 0.4 |

#### Sediment organic matter

Percentage of sediment organic matter was significantly greater in Mtoni (6.5%) than in Dege (5.1%) (Two-sample t test: t = 2.201, df = 77, p = 0.0307) at probability of 0.05.

#### Percent saturation capacity

Percent saturation capacity was extremely significantly greater in Mtoni (47.6%) than in Dege (31.1%) (Two-sample t-test: t = 5.265, df = 77, p = 1.23 x 10<sup>-6</sup>) at probability of 0.05.

#### Sediment interstitial salinity

Sediment interstitial salinity was very significantly greater in Mtoni (34.2%) than Dege (32.4%) (Two-sample t-test: t = 3.208, df = 77, p = 0.0019) at probability of 0.05.

# Correlations between mangrove vegetation characteristics and selected abiotic factors

At Mtoni, both basal area and mangrove density were significantly negatively correlated with percent organic matter (Table 3). While there was a significant negative correlation between basal area and salinity at Mtoni, the same parameters were very significantly positively correlated at Dege (Table 4). Organic matter and percent saturation capacity were very significantly positively correlated at both Mtoni and Dege. Mangrove density and mangrove basal area were also significantly positively correlated at Mtoni.

|                       | Mangrove    | Mangrove  | % Saturation | % Organic | Salinity |
|-----------------------|-------------|-----------|--------------|-----------|----------|
|                       | basal area  | density   | capacity     | matter    |          |
| Mangrove basal area   | 1.00        |           |              |           |          |
| Monanova donaity      | 0.664       | 1.00      |              |           |          |
| Mangrove density      | (<0.001) ** |           |              |           |          |
| % Saturation conscitu | 0.300       | -0.246    | 1.00         |           |          |
| % Saturation capacity | (0.101)     | (0.182)   |              |           |          |
| % Organia matter      | -0.437      | -0.358    | 0.877        | 1.00      |          |
| % Organic matter      | (0.014) *   | (0.048) * | (<0.001) **  |           |          |
| Solinita              | -0.369      | -0.195    | 0.177        | 0.202     | 1.00     |
| Samily                | (0.041) *   | (0.292)   | (0.339)      | (0.276)   |          |

Table 3. Correlations between mangrove basal area and mangrove density, and selected abiotic factors at Mtoni forest along the coast of Dar es Salaam

\* = significantly correlated, \*\* = very significantly correlated. Values shown represent Spearman's rank correlation coefficients, with the probability of a Type I error in parentheses. n = 31

 Table 4. Correlations between mangrove basal area and mangrove density, and selected abiotic factors at Dege forest along the coast of Dar es Salaam

|   | Mangrove<br>basal area | Mangrove<br>density | % Saturation capacity | % Organic matter | Salinity |
|---|------------------------|---------------------|-----------------------|------------------|----------|
| Mangrove basal area                       | 1.00                   |                     |                       |                  |          |
| Mangrove density<br>% Saturation capacity | 0.250                  | 1.00                |                       |                  |          |
|   | (0.90)                 | 0.220               | 1.00                  |                  |          |
|   | (0.792)                | (0.23)              | 1.00                  |                  |          |
| % Organic matter                          | -0.068                 | -0.221              | 0.856                 | 1.00             |          |
|   | (0.651)                | (0.135)             | (<0.001) **           |                  |          |
| Salinity                                  | 0.464                  | -0.150              | 0.332                 | 0.178            | 1.00     |
|   | (<0.001) **            | (0.313)             | (0.322)               | (0.230)          |          |

\*\* = very significantly correlated. Values shown represent Spearman's rank correlation coefficients, with the probability of a Type I error in parentheses. n = 48.

# DISCUSSION

#### Mangrove vegetation characteristics

The findings showed that mangrove species diversity and richness were greater in Mtoni than in Dege. This is contrary to what was expected according to Hypotheses No. 1. This may be due to the differences in the physical characteristics of the substrate. Dege has primarily rocky/muddy substrate, which is fairly uniform throughout the site. On the other hand, the substrate in Mtoni ranges from sandy in the upper zone to sandy/muddy in the middle zone and muddy in the lower zone. Due to the greater variation in the substrate, greater species diversity can be supported. In particular, the sandy substrate in the upper zone supports the additional species found, i.e., Lumnitzera racemosa.

Finding for this study found that there were 7 mangrove species in Mtoni, with a species

diversity index of 0.61, which are both higher than previous studies by Akwilapo (2001) and Mgaya *et al.* (2004) who found 3 mangrove species with a species diversity index of 0.42; and 6 mangrove species with a species diversity index of 0.65 respectively. The greater number of species found in this study may be due to the greater area covered for mangrove assessment. Results for Dege also differ from what was reported by Sallema (2003) who found a total of 5 mangrove species.

The findings, that the mangrove basal area was higher in Dege than Mtoni, verify Hypothesis No 1. The main reason is the high rate of mangrove exploitation in Mtoni, where it was found to be 39 stumps/100-m<sup>2</sup> plot in comparison with Dege where it was 26 stumps/100-m<sup>2</sup> plot. Mangrove users tend to harvest the larger trees, leaving the smaller ones which contribute little to the basal area. Thus, in Dege, there is a greater abundance of larger trees remaining. An important factor leading to this difference is the fact that Mtoni is closer to Dar-Es-Salaam city center, and thus there is greater demand for mangrove products.

Comparing the results of this study with those of previous studies shows that there has been a drastic decrease in the basal area over time at Mtoni. Akwilapo (2001) found a stand basal area of 600 cm<sup>2</sup>/25-m<sup>2</sup> plot and Mgaya et al. (2004) found a basal area of 446  $cm^2/25-m^2$  plot, while this study shows a basal area if converted to the same units, of 139  $\text{cm}^2/25\text{-m}^2$  plot. This suggests that mangrove cutting pressure has been increasing with time. The basal areas obtained in this study for Mtoni and Dege  $(252 \text{ cm}^2/25 \text{-m}^2 \text{ plot}, \text{ if converted to the same})$ units) are considerably lower than the values found in other mangrove forests in Tanzania that are located far from urban settlements, such as  $1261 \text{ cm}^2/25\text{-m}^2$  plot reported for the Rufiji estuary (Wagner et al. 2003) and 1015  $cm^2/25-m^2$  plot reported for the Ruvuma estuary (Wagner et al. 2004). However, the basal areas for Mtoni and Dege are higher than those reported by Wagner (2005) for Mbweni and Kunduchi (85 and 64 cm<sup>2</sup>/25 $m^2$  plots, respectively).

The fact that *Ceriops tagal* was dominant in both sites (with respect to both trees and saplings), may be due to preferential cutting of certain popular mangrove species, such as *Rhizophora mucronata* (used for building poles), as well as *Avicennia marina* and *Sonneratia alba* (used for firewood). *Ceriops tagal*, being short and bush-like in structure, is of little value to the people.

The dominance of *Ceriops tagal* corresponds with those of Masoud and Wild (2004) who noted that *Ceriops tagal* and *Rhizophora mucronate* were dominant species in Zanzibar, followed by *Bruguiera gymnorrhiza* and *Xylocarpus granatum*. The findings also relate to that of Mohamed (2004) who noted that the dominant species in Chake-Chake Bay are *Rhizophora mucronata* and *Ceriops tagal*. However, the findings of this study differ from those reported by Mgaya *et al.* (2004) that *Avicennia marina* had the highest density, followed by *Bruguiera gymnorrhiza* and *Ceriops tagal* in Mtoni. This is probably due to the fact that their study was conducted in only some areas where *Avicennia marina* was dominant; whereas, this study covered the whole area of Mtoni forest.

The higher regeneration capacity (as indicated by the density of seedlings) at Dege than Mtoni is probably related to the higher basal area and density of mangroves in the former site. Overexploitation of mature trees in Mtoni could have resulted in the absence of enough seedlings. It has been reported by Mabula (2017) that low regeneration can be caused by a lack of enough mature trees, absence of seedlings, seed predation, and strong tidal waves. Akwilapo (2001) and Wagner (2007) noted that the absence of *Sonneratia alba* seedlings in Mtoni was caused by fishermen who were dragging seine nets under the tree canopy.

The extremely high abundance of seedlings of *Ceriops tagal* in Dege is related to the fact that this species had by far the highest density of trees in that site. In Mtoni, there were high densities of *Avicennia marina*, *Ceriops tagal*, and *Bruguiera gymnorrhiza* seedlings. This could again, be due to the high tree densities of these species.

# Human pressures on mangroves

The finding showed that mangrove stump density was higher in Mtoni than in Dege. This verifies Hypothesis No. 2. The reason could be due to the greater mangrove Mtoni and subsequent harvesting in degradation compared to Dege. It is likely due to the location of Mtoni which is near larger human populations, that puts higher exploitative pressure on the forest products, such as firewood, charcoal, and building poles. In both study sites, the Ceriops tagal had the highest stump density. This might be due to its high abundance and small size. Its small size in terms of diameter could facilitate its removal even by using a small hand tool. Thus, it has higher availability,



even though it may not be the preferred species. Masoud and Wild (2004) reported Ceriops tagal to be the most exploited species in Zanzibar. Mfaume (2015) further notes that the stunted growth of Ceriops tagal is caused by both edaphic factors and anthropogenic overexploitation. Other species that followed for being exploited in both Mtoni and Dege are Rhizophora and Bruguiera. This was attributed to the fact that they are preferred for building poles and other construction works due to their good growth form and they usually attain intermediate size. However, Khamis et al. (2017) give general highlight that population increase, urbanisation, and emerging coastal activities confront mangrove forests in Tanzania.

# Correlations between mangrove vegetation characteristics and selected abiotic factors

Even though more abundant mangrove growth increases organic matter, mangrove abundance (basal area and density) and organic matter content of the substrate were significantly negatively correlated in Mtoni, which is contrary to what was expected according to Hypothesis No. 3. This might have been due to the fact that, in some plots, there was a large number of small trees (especially Ceriops tagal), which contribute less to leaf litter resulting in little soil organic matter, while in other plots, there were few larger trees at a low density that contribute more leaf litter. Akwilapo (2003) reported Mtoni to have a high abundance of polychaetes due to the sewage input. Polychaetes act as shredders (Dean2008) and play a significant role in the decomposition of mangrove litter; therefore, reducing the residence time of organic matter (Alongi et al. 2005).

In Dege, there was a very significant positive correlation between basal area and salinity while the same parameters were significantly negatively correlated in Mtoni. It has been explained by Alongi (2008) that mangroves, especially *Avicenia marina*, have the ability to tolerate a wide range of environmental conditions such as salinity and air temperature, which allows them to thrive in frequently inundated areas both and infrequently flooded upper intertidal areas. A combination of other influences such as biotic factors may have been the cause for the significant negative correlation between basal area and salinity in Mtoni. This finding corresponds to the report by Ouyang and Guo (2020) and Satyanarayana et al. (2010) that there is a significant negative correlation between salinity and both mangrove density and basal area.

Organic matter and percent saturation capacity were very significant positive correlated in both Mtoni and Dege. Mamidala *et al.* (2022) and Yan *et al.* (2021) have also shown that there is a significant positive correlation between organic matter and percent saturation capacity. This can be explained by the fact that organic matter has the ability to conserve moisture for a longer period of time. Mangrove density and mangrove basal area were also significantly positive correlated in Mtoni. This situation might have been attributed to the presence of few seedlings and saplings which in turn contribute less to the basal area.

# CONCLUSIONS

The findings from this study showed Mtoni to have higher mangrove species diversity than Dege. 7 mangrove species were observed in Mtoni and 6 in Dege. Dege had significantly higher mangrove basal area and density than Mtoni, undoubtedly due to lesser disturbance by human activities in the former. Regeneration capacity, as indicated by the density of seedlings, was greater in Dege than in Mtoni, which can be attributed to the higher abundance of mature trees in Dege compared to Mtoni. Mtoni had a significantly higher density of stumps than Dege. This indicates that there is greater degradation at Mtoni due to its close proximity to the Dar es Salaam city center, accompanying demands with its for firewood, charcoal, and building poles as well as the discharge of domestic and industrial pollutants. It is recommended that deliberate efforts to manage these mangrove ecosystems be devoted to mere conservation initiatives being undertaken for the Dege and active restoration for the Mtoni. Participation of the adjacent communities is crucial in ensuring sustainable management of these mangrove ecosystems. Awareness creation to the community, planners, and policymakers should also be extended for the safety of mangrove ecosystems.

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# REFERENCES

- Akwilapo, F.D. 2001. Distribution and abundance of mangrove benthic macrofauna in mangrove ecosystem showing different levels of anthropogenic degradation. M.Sc. Thesis, University of Dar es Salaam.
- Alongi, D.M. 2008. Mangrove forests: Resilience, protection from tsunamis and responses to global climate change. Estuarine, Coastal and Shelf Science 76:1-13.
- Alongi, D.M., Ramanathan, A., Kannan, L., Tirendi, F., Trott, L. & Bala Krishna, P.M. 2005. Influence of human disturbance on benthic microbial metabolism in the Pichavarum mangroves, Vellare-Coleroon estuarine complex, India. Marine Biology 147: 1033-1044.
- Basha, S.K. 2018. An overview on global mangroves distribution. http://nopr.niscpr.res.in/bitstream/123

456789/44254/1/IJMS%2047%284% 29%20766-772.pdf.

- Branoff, B.L. 2017. Quantifying the influence of urban land use on mangrove biology and ecology: A meta-analysis. Global ecology and biogeography 26(11): 1339-1356.
- Brown, I., Mwansasu, S. & Westerberg, L.O. 2016. L-band polarimetric target decomposition of mangroves of the Rufiji Delta, Tanzania. Remote Sensing 8(2): 140. file:///C:/Users/USER/Downloads/rem otesensing-08-00140-v2.pdf.
- Dean, H.K. 2008. The use of polychaetes (Annelida) as indicator species of marine pollution: a review. Journal of Tropical Biology 56(4): 11-38.
- Diniz, U.M., de Lima N.T., Mello, M.A.R. & Machado, I.C. 2022. Few plants and one dominant fly shape a unique pollination network in a Neotropical mangrove. Aquatic Botany180: 103526.
- English, S., Wilkinson, C. & Baker, V. 1997. Survey Manual for Tropical Marine Resources, 2<sup>nd</sup> Edition. Australian Institute of Marine Science, Townsville. 390 pp.
- Feka, Z.N. & Morrison, I. 2017. Managing mangroves for coastal ecosystems change: a decade and beyond of conservation experiences and lessons for and from west-central Africa. Journal of Ecology and The Natural Environment 9(6): 99-123.
- Friess, D.A., Rogers, K., Lovelock, C.E., Krauss, K.W., Hamilton, S.E., Lee, S.Y. & Shi, S. 2019. The state of the world's mangrove forests: past, present, and future. Annual Review of Environmental Resources 44(1): 89-115.
- Khamis, Z.A., Kalliola, R. & Kayhko, N. 2017. Geographical characterization of the Zanzibar coastal zone and its management perspectives. Ocean and Coastal Management 149: 116-134.

- Kristensen, E., Mangion, P., Tang, M., Flindt, M.R., Holmer, M. & Ulomi, S. 2011. Microbial carbon oxidation rates and pathways in sediments of two Tanzanian mangrove forests. Biogeochemistry 103(1): 143-158.
- Linneweber, V. 2013. Mangrove ecosystems: function and management. Springer Science & Business Media. http://10.1007/978-3-662-04713-2
- Lyimo, B.M. 2006. Energy and Sustainable Development in Tanzania. HELIO International Energy Watch, Dar es Salaam (8): 3-8.
- Mabula, M.J., Mangora, M.M. & Muhando, C.A. 2017. Peri-urban mangroves of Dar es Salaam-Tanzania are highly vulnerable to anthropogenic pressures. Advances in Ecological and Environmental Research: 141-172. https://www.researchgate.net/profile/ Makemie-Mabula/publication/315652258\_
- Malik, A., Fensholt, R. & Mertz, O. 2015. Mangrove exploitation effects on biodiversity and ecosystem services. Biodiversity and Conservation 24(14): 3543-3557.
- Mamidala, H.P., Ganguly, D., Ramachandran, P., Reddy, Y., Selvam, A.P., Singh, G. & Ramachandran, R.
  2022. Distribution and dynamics of particulate organic matter in Indian mangroves during dry period. Environmental Science and Pollution Research: 1-12.
- Manara, M. 2020. Land tenure formalization in Dar es Salaam: institutional transition through endogenous social interactions (Doctoral dissertation, The London School of Economics and Political Science (LSE). http://etheses.lse.ac.uk/4303/
- Mangora, M.M., Lugendo, B.R., Shalli, M.S. & Semesi, S. 2016. Mangroves of Tanzania. Mangroves of the Western Indian Ocean: status and management: 33-49.

- Maseta, G.J., Mwansasu, S. & Njana, M.A. 2021. Carbon dynamics and sequestration by urban mangrove forests of Dar es Salaam, Tanzania. Western Indian Ocean Journal of Marine Science 20(2): 11-23.
- Masoud, T.S. & Wild, R.G. 2004. Sustainable use and conservation management of mangroves in Zanzibar, Tanzania. Mangrove management and conservation: present and future: 280-293.
- Mfaume, A.H. 2015. The extent of mangrove deforestation and environmental impacts in Menai Bay, Zanzibar, Doctoral dissertation. University of Dodoma. http://41.78.64.25/handle/20.500.1266 1/610.
- Mgaya, Y.D., Machiwa. J.F., Lugomela, C., Lyimo, T., Joseph, C., Muzuka, A., Shaghude, Y., Mvungi, A., Mapunda, Mwanuzi. F.L. 2004. B. & Environmental studies of Mtoni mangrove ecosystem, Dar es Salaam. Final report submitted to the Faculty of Science, University of Dar es Salaam, Dar es Salaam: 123 pp.
- Mihale, M.J., Tungaraza, C., Baeyens, W. & Brion, N. 2021. Distribution and sources of carbon, nitrogen and their isotopic compositions in tropical estuarine sediments of Mtoni, Tanzania. Ocean Science Journal 56(3): 241-255.
- Mohamed, S.M. 2004. Investigation on species diversity and the extent of deforestation in Chake-Chake Bay mangrove ecosystem, Pemba Island, Tanzania, M.Sc. thesis, University of Dar es Salaam, Dar es Salaam.
- Moshi, I., Msuya, I.R. & Todd, G. 2018. Tanzania: National urban policies and city profiles for Dar es Salaam and Ifakara.

https://eprints.gla.ac.uk/208151/

Mtanga, A. & Machiwa, J. 2008. Assessment of heavy metal pollution in sediment and polychaete worms from the Mzinga Creek and Ras Dege mangrove ecosystems, Dar es Salaam, Tanzania. Western Indian Ocean Journal of Marine Science 6(2).

- Ndetto, E.L. & Matzarakis, A. 2013. Basic analysis of climate and urban bioclimate of Dar es Salaam, Tanzania. Theoretical and Applied Climatology 114(1): 213-26.
- Nolan, K.A. & Callahan, J.E. 2006. Beachcomber biology: The Shannon-Weiner species diversity index. InProc. Workshop ABLE (27): 34-338.
- Ouyang, X., & Guo, F. 2020. Patterns of Mangrove Productivity and Support for Marine Fauna. Handbook of Halophytes: From Molecules to Ecosystems towards Biosaline Agriculture, 1-20.
- Sallema, A.E. 2003. Mangrove species distribution, abundance, and restoration potential in Dege mangrove ecosystem. M. Sc. Dissertation, University of Dar es Salaam, Dar es Salaam.
- Satyanarayana, B., Idris, I.F., Mohamad, K.A., Husain, M.L., Shazili, N.A. & Dahdouh-Guebas, F. 2010. Mangrove species distribution and abundance in relation to local environmental settings: a case-study at Tumpat, Delta, Kelantan east coast of peninsular Malaysia. https://doi.org/10.1515/BOT.2010.006
- Thomas, N., Bunting, P., Lucas, R., Hardy, A., Rosenqvist, A. & Fatoyinbo, T. 2018. Mapping mangrove extent and change: a globally applicable approach. Remote Sensing 10(9): 1466.
- Todd, G., Msuya, I., Levira F. & Moshi, I. 2019. City profile: Dar es Salaam, Tanzania. Environment and Urbanization ASIA 10(2): 193-215.
- UNEP. 2014. Mangroves: Biodiversity a-z. UNEP-WCMC. http://biodiversityaz.org/content/mangrove--2.pdf.

- Wagner, G.M. 2005. "Participatory monitoring of changes in coastal and marine biodiversity", Indian Journal of Marine Sciences 34(1): 136-146.
- Wagner, G.M. 2007. The Dar es Salaam Seascape: A Case Study of an Environmental Management 'Hotspot'. file:///C:/Users/USER/Downloads/482 29-Article%20Text-61900-1-10-20091126%20(2).pdf.
- Wagner, G.M., Akwilapo, F.D., Mrosso, S., Ulomi, S. & Masinde, R. 2004. Assessment of marine biodiversity, ecosystem health and resource status in mangrove forests in Mnazi Bay-Ruvuma Estuary Marine Park (MBREMP). Final report submitted to IUCN-EARO, the World Conservation Union Eastern Africa Regional Office. 106 pp. https://portals.iucn.org/library/efiles/d ocuments/2004-118.pdf.
- Wagner, G.M., Makota, V. & Sallema, R. 2003. "Mangrove forests", in Tanzania State of the Coast Report (2003), The National ICM Strategy and Prospects for Poverty Reduction. Tanzania Coastal Management Partnership (TCMP), Dar-Es-Salaam: 5-11.
- Webber, M., Calumpong, H., Ferreira, F., Granek, E., Green, S., Ruwa, R. & Soares, M. 2016. Mangroves. The First Global Integrated Marine Assessment: World Ocean Assessment I, 877-886. https://www.researchgate.net/profile/F lavia-Lucena.
- Yan, L., Xie, X., Peng, K., Wang, N., Zhang, Y., Deng, Y. & Zhang, Y. 2021. Sources and compositional characterization of chromophoric dissolved organic matter in a Hainan tropical mangrove-estuary. Journal of Hydrology 600: 126572.
- Zar, J.H. 1999. Biostatistical analysis, 4<sup>th</sup> Edition. Simon and Aviacon Company. Upper Saddle River, New Jersey: 663 pp.